Historical Changes in Sediment and Phosphorus Loading to the Upper Mississippi River:

Mass-balance Reconstructions from the Sediments of Lake Pepin

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by

Daniel R. Engstrom, Ph.D. and James E. Almendinger, Ph.D.



St. Croix Watershed Research Station Science Museum of Minnesota Marine on St. Croix, MN 55047

with a contribution on Sediment Apportionment

by

David W. Kelley and Edward A. Nater, Ph.D.

Department of Soil, Water & Climate University of Minnesota St. Paul, MN 55108

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ABSTRACT

Long-term changes in sediment and phosphorus loading to the upper Mississippi River were quantified from an array of 25 sediment cores from Lake Pepin, a large natural impoundment downstream of the Twin Cities metropolitan area. Cores were dated and stratigraphically correlated using 210 Pb, 137 Cs, 14 C, magnetic susceptibility, pollen analysis, and loss-on-ignition. All cores show a dramatic increase in sediment accumulation beginning with European settlement in 1830. Accumulation rates are highest and show the greatest post-settlement increases in the upper end of the lake. Present-day sediment-phosphorus concentrations are roughly twice those of pre-settlement times, and the Fe/Albound fraction makes up a greater portion of the total. Diatom assemblages record a marked increase in nutrient availability over the last 200 years, changing from clear-water benthic forms and mesotrophic planktonic taxa in pre-settlement times to exclusively planktonic assemblages characteristic of highly eutrophic conditions today. Lake-water total-phosphorus concentrations (total-P), estimated by weighted averaging regression and calibration, increased from 50 to 200 μ g l⁻¹ during this period.

Sediment loading to Lake Pepin from the Mississippi River has increased by an order of magnitude since 1830. Modern fluxes are about 900,000 metric tons annually, and are more than 80% detrital mineral matter. About 17% of the lake's volume in 1830 has been replaced by sediment, and at current accumulation rates the remainder will be filled in another 340 years. Phosphorus accumulation in Lake Pepin sediments has increased 15-fold since 1830, rising from 60 to 900 metric tons annually. This rise represents a seven-fold increase in phosphorus loading from the Mississippi River coupled with more efficient retention of phosphorus inflows by bottom sediments. More efficient trapping of phosphorus in Lake Pepin over the last century resulted from higher rates of sediment burial. The most dramatic changes in nutrient and sediment inputs to Lake Pepin have occurred since 1940, although gradual increases began shortly following European settlement. Sediment accumulation rates rose sharply between 1940 and 1970 and then leveled off, while phosphorus inflows record their largest increases after 1970.

INTRODUCTION

The Mississippi is the largest and economically most important river in North America. It drains 40% of the conterminous United States and carries each year an estimated 80,000 metric tons (t) of phosphorus, 400,000 t of nitrate and 70,000,000 t of suspended sediment to the Gulf of Mexico (Smith et al., 1994). These loads have increased within recent decades resulting in eutrophication of the continental shelf and possible areal expansion of summer hypoxia in the northern Gulf (Rabalais et al., 1994; Turner & Rabalais, 1994; Justic et al., 1995). Impaired water quality is an equally serious concern for the upper Mississippi River, where a combination point and nonpoint sources of phosphorus have created highly eutrophic conditions that in 1988 caused severe algal blooms and fish kills downstream of the Minneapolis-St. Paul (Twin Cities) metropolitan area (Heiskary & Vavricka, 1993; MWCC, 1993d). These concerns are heavily focused on Lake Pepin, a large natural impoundment of the Mississippi below the Twin Cities and an important recreational and commercial resource for the region. Lake Pepin is currently the subject of a major inter-agency water-quality study designed to quantify impacts of point and nonpoint phosphorus loads within the basin (MWCC, 1993a). A key element of that study, on which we report here, is the historic assessment of long-term changes in sediment and phosphorus loading to the Mississippi above Lake Pepin. This information is critical to understanding the extent of the eutrophication problem compared to natural background conditions and for determining present rates of change and future projections for the river (Smol, 1992).

Historic water quality data, even for a river of the Mississippi's stature, are of relatively short duration. Reliable estimates of nutrient loads to Lake Pepin, which go back to the early 1970s, suggest a steady increase in phosphorus inputs (adjusted for flow variations) over the last two decades (Heiskary & Vavricka, 1993). Earlier records are sparse to nonexistent, and long-term change instead must be inferred from sediments accumulated in depositional environments along the Mississippi's course. Rising concentrations of biogenic (diatom) silica in sediment cores from the Gulf coast have been cited as evidence for 20th century eutrophication by the Mississippi River (Turner & Rabalais, 1994), but such information is neither quantitative nor does it apply to specific upstream reaches. With few exceptions, there is little historical evidence for eutrophication of the world's major rivers, though most are considered to be highly impacted by human activities in their watersheds. The major problem is that most rivers do not possess along their course depositional basins where sediments accumulate conformably, and in this respect Lake Pepin is nearly unique. In this investigation we compile, through analysis of multiple sediment cores, a quantitative estimate of sediment and phosphorus loading to the upper Mississippi River over the last several hundred years.

Lake sediment cores have been used in numerous studies to produce qualitative estimates of past trophic state, the most common proxies being the accumulation rate of phosphorus (P) and the composition of fossil diatom assemblages. Both approaches have become more quantitative in recent years, with whole-basin P accumulation now calculated from multiple sediment cores (Evans & Rigler, 1980; Moss, 1980; Brezonik & Engstrom, 1998), and lakewater P concentrations numerically reconstructed by multivariate transfer-function techniques (Hall & Smol, 1992; Anderson & Rippey, 1994; Bennion, 1994). In an elegant synthesis of these methods, Rippey and Anderson (1996) demonstrated that historical P loadings to lakes could be rigorously determined from a mass-balance sum of P sedimentation and outflow (the latter calculated from historic flows and diatom-inferred lake-water total-P). The major advantage of this mass-balance approach (over diatoms or sediment-P alone) is that P sedimentation is strongly influenced by redox changes that accompany eutrophication (Engstrom & Wright, 1984; Schelske et al., 1988) as well as changes in hydraulic residence time which are pronounced in riverine lakes like Pepin. Hence, we adopt in this study a combination of diatom-inference and whole-lake sedimentation methods to provide the first quantitative account of human alteration of phosphorus and sediment loads to one of the world's great rivers.

STUDY AREA

Lake Pepin is a natural floodplain lake on the upper Mississippi River located about 80 km downstream from the Twin Cities metropolitan area of Minnesota. The lake formed about 10,000 years ago behind an alluvial fan of the Chippewa River, which dammed the Mississippi after outflow from Glacial Lake Agassiz was diverted northward and ceased to scour sediments deposited by the Mississippi's tributaries (Wright *et al.*, 1998). Although much larger in the distant past, Lake Pepin today has a length of 34 km, a mean depth of 5.4 meters, and a surface area of 103 km² (Fig. 1). Mean annual flows of the Mississippi River above Lake Pepin are about 600 m³ s⁻¹ (Stark *et al.*, 1996), and residence time in the lake is relatively short (6-47 days) (Heiskary & Vavricka, 1993). Water quality in Lake Pepin today is generally poor, reflecting the combined impacts of urban and agricultural developments upstream. Mean summer total-P concentrations at the inflow to Lake Pepin averaged 215 μ g l⁻¹ and total suspended solids concentrations averaged 41 mg l⁻¹ between 1977 and 1991 (Heiskary & Vavricka, 1993).

The Mississippi River above Lake Pepin drains about 122,000 km², mostly in Minnesota and adjacent Wisconsin, through many small tributaries and two major rivers, the Minnesota

and St. Croix (Fig. 1). These two tributaries each provide about one-fourth of the mean annual flow entering Lake Pepin, although the Minnesota's watershed is more than twice that of the St. Croix (Stark *et al.*, 1996). Precipitation and runoff are substantially higher in the eastern regions drained by the St. Croix. The watershed of the Minnesota is 85 to 90% agricultural, (predominately row crops), and presently contributes 75% of the suspended solids and 32% of the phosphorus loads to the upper Mississippi River (UMR) above Lake Pepin in an average flow year (MWCC, 1993b; MWCC, 1994). The St. Croix's watershed, on the other hand, is largely forested and provides only about 4% of the average phosphorus inputs to the UMR above Lake Pepin. The watershed of the Mississippi main stem is forested in the north, grading to agricultural (pasture and dairy) in the south; it provides about 26% of the P loads to the UMR. The Minneapolis-St. Paul urban area (1990 population of 2.3 million) contributes about 24% of the remaining P loads (under average flow conditions) through its waste-treatment facilities, while other point-source discharges and smaller tributaries make up the difference (MWCC, 1993b).

METHODS

Sediment Coring: A total of 25 sediment cores were collected (September, 1995 and June and October, 1996) along five shore-to-shore transects positioned perpendicular to the flow axis of Lake Pepin (Fig. 1). The transects were distributed at roughly equal distances from upper to lower reaches of the lake, and five cores were taken along each transect. Cores are designated here by transect number (I to V from upper to lower lake) and position along the transect (1 to 5 from the Minnesota to the Wisconsin side of the Lake); for example core V.1 is from the lowermost transect closest to the Minnesota side. Core locations were recorded in the field by differential GPS (global positioning system). Additional reconnaissance cores were collected to help delineate the depositional limit of fine-grained sediments in Lake Pepin; these cores were discarded in the field.

Sediments were collected during the open-water season using piston corers operated from the lake surface by rigid drive-rods. A surface corer equipped with a 7-cm diameter polycarbonate core barrel was used to collect a continuous 2-m section of the upper sediments at all coring sites (Wright, 1991). This device recovers the watery, uncompacted sediment surface as well as deeper strata without disturbance or displacement (coreshortening) (Blomqvist, 1985; Blomqvist, 1991). Additional 1-m overlapping drives were taken below the surface core-section with a square-rod Livingstone corer (Wright *et al.*, 1983) at those sites where very rapid sediment accumulation was anticipated (transects I-III); total core length at these locations was 3.5 to 4 m. The surface core-sections were held upright until the upper 20-24 cm of soft sediment could be extruded vertically (at 2-cm increments) into polypropylene collection jars on shore. The remaining section of stiff sediment was capped in the core tube and transported back to the lab in horizontal position. The Livingstone sections were extruded on site and wrapped in polyethylene film (foodwrap) and aluminum foil. The cores were stored at 4° C until further processing.

<u>Seismic Survey</u>: To specifically delineate the depositional limit of fine-grained sediments in Lake Pepin, high resolution single channel seismic reflection data were acquired along 26 shore-to-shore transects (Fig. 1) using an ORE Geopulse boomer system. The data were digitally recorded on a Triton ISIS acquisition system. Differential GPS was used for navigation. Post-acquisition processing was performed on the data with the SU seismic processing package distributed by the Colorado School of Mines. The applied processing flow consisted of: editing of bad traces, geometry and gain adjustment, filtering, and deconvolution for signal enhancement.

<u>Mapping</u>: Core locations, seismic-survey routes, bathymetric data, and geographic reference points for Lake Pepin were precisely mapped using ArcInfo GIS technology.

Rasterized bathymetric data, provided by the USGS Environmental Management Technical Center, Onalaska, WI, were contoured at 1-m intervals and used together with reconnaissance coring and seismic results to delineate depositional regions for fine-grained sediment. Throughout much of Lake Pepin the depositional limit (above which fine-grained sediments do not accumulate) corresponds closely to a sharp bathymetric scarp separating shallow nearshore areas from a generally flat central basin. Hence, the depositional limit could be mapped along this scarp everywhere except in the shallow upper end of the lake where the limit was approximated at the middle of the 1-m contour. The total depositional area of Lake Pepin was subdivided into five depositional regions, one for each transect of cores, by the perpendicular bisector between adjacent transects (Fig. 1).

Magnetic Susceptibility: For all 25 cores, whole-core magnetic susceptibility measurements were made at 2-cm intervals using a Bartington MS2 core logging sensor and automated track. These measurements provide a non-destructive determination of the concentration of fine-grained ferromagnetic minerals within the sediments. Cores were measured in 1-meter sections, and magnetic profiles were spliced between sections according to sediment depth. In some cores it was necessary to make depth adjustments of a few centimeters between overlapping sections to align magnetic features. Susceptibility readings for the smaller diameter Livingstone core-sections were scaled to those of the surface core-sections before splicing records.

Core Sampling: Following measurement of whole-core magnetic susceptibility, ten of the 25 cores were selected for detailed stratigraphic analysis (two per transect) and were subsectioned into 4-cm increments. Each core-section was scraped with a stainless steel spatula to remove the outside smear and subsectioned with a thin wire. In most cases, only the surface core-section was completely subsampled; the 1-m Livingstone core-sections were stored intact and only selected 4-cm increments were removed for analysis. An additional 10 cores were chosen for "top/bottom" stratigraphic analysis in which constituents were analyzed from just two time-stratigraphic intervals. Core-top samples ranged in length from 12 to 24 cm (depending on sedimentation rate) and represented on average the last 7 years of sediment deposition. Core-bottom samples, 4 to 15 cm in length, were taken from strata deposited just prior to European settlement and represented an average accumulation of 60 years.

<u>Loss-on-Ignition</u>: Dry-density (dry mass per volume of fresh sediment), water content, organic content, and carbonate content of Pepin sediments were determined by standard loss-on-ignition techniques (Dean, 1974). Sediment samples of 1-2 g were dried overnight at 100° C and ignited at 550° and 1000° C for 1 hr each. Mass measurements were made of the wet samples and after each heating on an electronic analytical balance. Dry density was calculated from water content and fixed densities for organic, carbonate, and inorganic fractions.

<u>Pollen Analysis</u>: Pollen samples from five cores were prepared according to procedures described by Faegri and Iversen (1975). A known quantity of *Eucalyptus* pollen was added to selected samples as a tracer to permit calculation of pollen concentration. Residues were mounted in silicon oil, and pollen identified under magnifications of 400X and 1000X. At least 300 terrestrial pollen grains were counted in each sample.

Radiocarbon Dating: Small pieces of terrestrial woody material were sieved from a selected depth interval in each of two cores. The selected intervals corresponded to magnetic features identifiable in many of the other cores. Samples were washed in deionized water, dried, and submitted to the Limnological Research Center of the University of Minnesota for graphite target preparation. The targets were analyzed by accelerator mass spectrometry (AMS) at the Center for Accelerator Mass Spectrometry of the Lawrence

Livermore National Laboratory. Radiocarbon dates were converted to calendar years AD using the CALIB program of Stuiver and Reimer (1993) and calibration data sets from Stuiver and Pearson (1993).

<u>Cesium-137 Dating</u>: Twenty sediment cores were analyzed for ¹³⁷Cs to identify sediments deposited during the 1963-64 peak in atmospheric nuclear testing. Freeze-dried samples (ca. 20 g) were measured for ¹³⁷Cs at 667 keV using a high-resolution germanium diode gamma detector and multichannel analyzer. Detector efficiency was around 2% as determined using a NIST-certified source with mineralogical composition similar to the samples (Rocky Flats soil / SRM4353). Accuracy and precision were better than 10% for a 12 hour measurement of concentrations of ~30 Bq/kg.

Lead-210 Dating: Ten sediment cores were analyzed for ²¹⁰Pb activity to determine age and sediment accumulation rates for the past 130 years. Lead-210 was measured at 17-24 depth intervals in each core through its grand-daughter product ²¹⁰Po, with ²⁰⁹Po added as an internal yield tracer. The polonium isotopes were distilled from 1.5 - 7.0 g dry sediment at 550° C following pretreatment with concentrated HCl and plated directly onto silver planchets from a 0.5 N HCl solution (Eakins & Morrison, 1978). Activity was measured for 1-3 x 10⁵ s with ion-implanted or Si-depleted surface barrier detectors and an Ortec alpha spectroscopy system. Unsupported ²¹⁰Pb was calculated by subtracting supported activity from the total activity measured at each level; supported ²¹⁰Pb was estimated from the asymptotic activity at depth (the mean of the lowermost samples in a core). Dates and sedimentation rates were determined according to the c.r.s. (constant rate of supply) model (Appleby & Oldfield, 1978) with confidence intervals calculated by first-order error analysis of counting uncertainty (Binford, 1990).

<u>Grain-Size Analysis</u>: Sediment grain size was measured in ten cores by laser analytical methods on a Lasentec Labtec-100 particle size analyzer. Samples were pretreated with 30% $\rm H_2O_2$ to oxidize organic matter and dispersed in 0.25% Calgon solution by shaking and ultrasonic mixing. Particles in each sample were assigned to 38 size classes (0.4 - 250 μ m) based on the average of 40 scans (four cycles of ten scans each) per sample, and mean grain size was calculated from this frequency distribution.

Source Apportionment: Sediment source-signatures were determined for the Upper Mississippi, Minnesota, and St. Croix river basins based on the elemental composition of heavy minerals in ten sediment grab-samples collected from the lower reaches of each river. Samples from ten Lake Pepin cores were then analyzed for their heavy-mineral elemental composition and matched statistically to their sources using a chemical mass-balance model. Sediment samples were treated to remove carbonates, organic matter, and iron oxides prior to particle size and density separation of the heavy silt fraction. Particle size separation was performed by sieving, and light and heavy silts were separated by heavy-liquid (sodium polytungstate) high-speed centrifugation. The heavy silt fraction was retained and analyzed because it typically has a much higher concentration of lanthanide series and rare-earth elements (REE) than does the light fraction. Samples were analyzed for 42 elements including major and trace rock elements and rare earth elements by Inductively Coupled Plasma-Mass Spectrometry (ICP-MS).

The U.S. Environmental Protection Agency's (USEPA) Chemical Mass Balance model (CMB version 8.0, 32 bit - EPA receptor model) was used to apportion lake sediments to their watershed sources. This model was designed for air resource management and is used routinely to apportion airborne particulates to their emission sources. The model requirements included the concentrations and standard deviations for the elemental analyses of the Lake Pepin cores (as outlined above) and of the three source rivers. Other than the transport medium, all other factors and requirements were the same as for modeling PM₁₀

ambient pollution concentrations. CMB uses an effective variance least squares estimation algorithm to fit known source elemental profiles to those of a receptor site. The least squares solution to a set of linear equations allows the model to express each receptor concentration of an element as a linear sum of products of source profile elements and source contributions. In this study, the model first established a source profile for the three rivers, then matched the relative contributions of these signature profiles to the elemental data from the Lake Pepin core samples. Performance measures are reported as r² and chi² values, with >0.8 and <4 values, respectively, considered a good fit between measured and calculated species concentrations.

Samples from 20 sediment cores were extracted for phosphorus Phosphorus Analysis: according to fractionation procedures adapted from Hieltjes and Lijklema (1980) and Plumb (1981). Total phosphorus was measured as the ortho-P extracted by block-digestion with concentrated H₂SO₄, K₂SO₄, and HgO. A second sediment aliquot was extracted in 1 M NH₄Cl for exchangeable-P (NH₄Cl-P), followed sequentially by 0.1 M sodium hydroxide for Fe/Al-bound phosphorus (NaOH-P). The sediment residue from the hydroxide extraction was further treated with 0.5 M HCl to determine carbonate-bound phosphorus (HCl-P). Finally the residual (organically-bound) phosphorus was estimated as the difference between total-P and the sum of the NH₄Cl, NaOH, and HCl extractions. Extractions were carried out on wet sediment samples, and water content was measured simultaneously on separate aliquots. Extracts were separated from sediment residue by centrifugation and membrane filtration. Phosphorus extracts were analyzed for orthophosphorus by the ascorbic acid method. Standard absorbance curves were constructed from an analytical P-standard in the appropriate matrix for each fraction. The reproducibility of sample replicates averaged 2.9% for total-P, 9.2% for NH₄Cl-P, 5.2% for NaOH-P, and 4.6% for HCl-P.

<u>Diatom Analysis</u>: Ten subsamples from core V.1 were analyzed for diatom microfossils. Normal cleaning procedures with H_2O_2 or sulfuric acid were not adequate for these sediments because of their high silt content. As an alternative method, uncleaned sediment samples were diluted with distilled water, settled, and dried overnight onto cover slips. The dried samples were then heated at high temperature to remove any organic matter. Coverslips were mounted on slides with Naphrax®. Diatom valves and chrysophyte cysts were counted in transects with a Leitz Ortholux microscope at 1200x magnification under oil immersion. A minimum of 500 diatom valves were counted in each sample except where diatom concentrations were very low. In these cases, a minimum of 400 valves or six transects were counted.

Lake-water total phosphorus (TP) was reconstructed from fossil diatom assemblages using diatom-phosphorus calibration models. Such models are developed using canonical correspondence analysis (CCA) and weighted averaging regression on the distribution of diatom species among lakes of known water chemistry (ter Braak & Prentice, 1988; Birks et al., 1990). All taxa composing more than 1% in any one sample were selected for TP reconstruction. Some taxa were further eliminated if TP optima were not available from the literature and they composed less than 5% in any one sample. Several published calibration data sets for TP reconstruction in lacustrine systems were tested on the Lake Pepin data, as no calibration data set is presently available for large rivers. Although Lake Pepin is essentially a riverine system, the flora is primarily composed of lacustrine diatoms (predominately Aulacoseira and Stephanodiscus species) that are well represented in most lake-based calibration data sets. The majority of species optima used for the Lake Pepin analysis were obtained from a calibration set of 31 southeast English lakes (Bennion, 1994) with TP values of 25-646 μ g l⁻¹. This set was chosen because it represents high-nutrient systems similar to Lake Pepin. Because some low-nutrient taxa present in Lake Pepin are absent or poorly represented in Bennion (1994), additional species optima were obtained

from Reavie *et al.* (1995). This second calibration set is based on 59 lakes in British Columbia with a TP range of 5-85 μ g l⁻¹.

RESULTS AND DISCUSSION

<u>Seismic Profiles</u>: The quality of the seismic reflection data was poor overall, and in some areas of the lake bottom gas bubbles in the sediment from decomposing organic matter severely attenuated the signal. Signal quality was especially poor in areas where the water depth was less than 4 m. In these situations the lateral offset of the source from the receiving hydrophone was greater than 2X water depth and little backscattered energy was recorded off the lake floor.

The seismic data show a parallel series of 3-4 strong horizontal reflectors at sediment depths of 0-2 meters throughout the flat mid-channel of the basin (Fig. 2). The reflectors appear at similar depths at both upper and lower ends of the lake, but they terminate abruptly on the steep bathymetric scarp on the sides of the channel. Unfiltered seismic data show a series of false "multiples" below the true reflectors, but no deeper features were detected anywhere in the lake, presumably because of rapid attenuation of acoustical energy by gas (methane) bubbles trapped in the sediment. The disappearance of the reflectors on the steep margins of the central basin indicates that fine-grained sediments are largely absent from these near-shore regions of the lake bottom (Fig. 2). Reconnaissance coring during the seismic survey obtained only coarse sand or fine gravel shoreward of this bathymetric discontinuity, confirming the seismic observations.

One of the seismic-survey routes passed directly over core-transect I (Fig. 1), allowing a direct comparison of sediment stratigraphy with the seismic data. The seismic profiles were aligned with core-tops at the uppermost (lake bottom) reflector and scaled to depth according to the estimated velocity of sound in saturated sediments. Comparisons were made with magnetic susceptibility and loss-on-ignition at four core-sites under the assumption that these sedimentological measures would most likely correspond to discontinuities in sediment density that might be captured by seismic reflection methods (Fig. 3). However, none of the seismic reflectors show any consistent alignment with stratigraphic features from the cores, implying that the seismic signals must derive from transient sedimentological changes such as sediment compaction, de-watering, or gasentrapment. Whatever the source for the seismic reflectors, they cannot be used to the map sedimentary facies revealed by the cores.

Magnetic Susceptibility: Whole-core magnetic susceptibility profiles show a striking pattern of peaks and troughs that can be readily traced among cores within each transect and, for the stronger features, from transect to transect throughout the lake basin (Fig. 4). The features were provisionally matched among profiles and numbered sequentially downcore. Subsequent dating analyses showed that the major magnetic features are time-synchronous and thus can be used to map time-stratigraphic units throughout the lake. Magnetic susceptibility, which is largely dependent on the concentration and size of magnetic mineral grains, is generally higher in the upper end of the lake and decreases downstream. Magnetic profiles in cores from the upper lake also show greater stratigraphic resolution than profiles from the lower lake, largely because of higher sediment accumulation rates in the upper lake. For example, magnetic feature (mf) 2 is a well defined trough in transects I and II but becomes a nearly imperceptible notch in transects IV and V. Likewise mf-4 and mf-6 are well separated peaks in transects I - III that become fused into a single peak in most of the transect V cores.

Magnetic feature 7 marks the initial rise in susceptibility from a low baseline and is the earliest stratigraphic evidence for European-settlement land-use changes in the upper

Mississippi watershed (see dating section). A sharp rise in magnetic susceptibility is commonly associated with increased soil erosion caused by forest clearance and tillage agriculture (Oldfield *et al.*, 1978; Dearing *et al.*, 1981; Oldfield *et al.*, 1989). Two small pre-settlement magnetic peaks (mf-7.5 and 8) are identified in a smaller subset of cores (Fig. 4). Although these features are relatively easy to trace within transects, their correlation between core transects requires additional evidence from pollen analysis. A positive match of mf-7.5 is possible because this feature is associated with an increase in "big woods" vegetation in all five cores counted for pollen. Magnetic feature 8 has no clear pollen marker, except that it precedes the "big woods" change, so its correlation among transects is more tentative.

Magnetic susceptibility was used as a reconnaissance tool to select which cores would be analyzed in stratigraphic detail (two from each transect) and which would be analyzed for top (modern) and bottom (pre-settlement) samples (Fig. 4). A few magnetic profiles (I.5, III.5, IV.1 and IV.5) were difficult to match with the others and were excluded from subsequent analysis. All such problematic cores were taken from sites very near the steep bathymetric scarp at the sides of the channel where depositional discontinuities such as slumping or scouring are most likely. These core sites were selected to explore sediment stratigraphy in "atypical" depositional environments and likely represent only a small proportion of the lake bottom.

Loss-on-Ignition (LOI): The percentages of organic matter, calcium carbonate, and inorganic matter in Lake Pepin sediments vary stratigraphically as well as spatially within the basin (Fig. 5). Organic matter averages 8.3% for all cores, but is lower at the upper end of the lake (7.3% in transect I) and increases progressively down lake (to 9.3% in transect IV). Carbonate content, on the other hand, shows the opposite pattern, decreasing from 12.8% in transect I to 8.3% in transect V. The inorganic component, largely comprised of clastic silicates, makes up the remaining 81% of the sediment and shows no longitudinal trend.

More striking, however, is the sharp up-core increase in organic matter (2-4%) and inorganic constituents (4-6%) and corresponding drop in carbonate (6-10%) that occurs in all 10 cores analyzed in stratigraphic detail (Fig. 5). This transition occurs at greater core depth in the upper lake (deeper than 2 meters in transect I) and at shallower depths in lower lake cores (about 1 meter in transect V). The organic-matter increase is most pronounced in transect I, but the inorganic rise is equally sharp in all parts of the basin. Organic matter remains elevated from this transition to the sediment surface, while inorganic content peaks and then declines (as carbonate increases) toward the tops of the cores. In each core the loss-on-ignition transition corresponds exactly to a point mid-depth between magnetic features 6 and 7 in the magnetic susceptibility profiles. The LOI change is also precisely correlated with pollen evidence for the onset of European-style agriculture and likely records increased soil erosion that accompanied widespread land clearance and watershed disturbance in the mid-1800s.

<u>Pollen Analysis</u>: Pollen profiles constructed for five of the ten detailed cores reveal a clear stratigraphic change from a pre-settlement assemblage dominated by pine (*Pinus*), oak (*Quercus*), birch (*Betula*), and mesic hardwoods (big woods) to one with dramatically increased percentages of weedy herbaceous taxa, particularly ragweed (*Ambrosia*) and chenopods (Chenopodiaceae), the first appearance of cereal pollen grains (corn, wheat, etc.), and a sharp reduction in pine pollen (Fig. 6). This transition represents the rapid spread of European-style agriculture in southern Minnesota and adjacent Wisconsin and the near-simultaneous logging of local pine forests along the lower St. Croix River. The exact timing of this "*Ambrosia* rise" is difficult to fix for a watershed the size of that contributing to Lake Pepin, but it is likely associated with the onset of large-scale

commercial agriculture, which began around the time of Minnesota statehood (1858). The *Ambrosia* rise coincides stratigraphically in each core with loss-on-ignition evidence for sharply increased erosion (see above), indicating that landscape disturbance had become fairly widespread by this time. However, both of these changes (pollen and LOI) are preceded by an earlier rise in magnetic susceptibility that likely records the earliest phases of European settlement and land-clearance in the Lake Pepin watershed (Fig. 4). The timing of this earlier magnetic marker is also uncertain, but a date of 1830, shortly following the establishment of Fort Snelling (at the confluence of the Minnesota and Mississippi rivers), is a reasonable first estimate. Both dates, 1860 for the Ambrosia rise and 1830 for the susceptibility increase, have an uncertainty of roughly ± 10 years.

The sediments of Lake Pepin record a more subtle change in pollen composition that appears about a meter below the Ambrosia rise and at the same depth as magnetic feature 7.5 in the five analyzed cores (Fig. 6). The most important signature of this shift is the increase in "big woods" pollen types (*Acer*, *Tilia*, *Ulmus*, *Fraxinus*, *Ostrya/Carpinus*, and *Carya*); ragweed and other herbaceous taxa decline at this transition, while birch, willow (*Salix*) and alder (*Alnus*) show less consistent changes. These palynological changes are clearest in cores II.3, III.2, and III.4, and less obvious in cores I.3 and V.1. This pollen horizon is identified with the regional expansion of "big woods" vegetation that accompanied the onset of cooler, wetter climatic conditions several hundred years before European settlement (Grimm, 1984). Although the timing of this vegetation change is somewhat uncertain, the pollen horizon itself provides a means of confirming stratigraphic correlation among cores in Lake Pepin. In particular it provides positive identification of mf-7.5 in susceptibility profiles among widely spaced core transects (see above).

Radiocarbon Dating: Two AMS ¹⁴C dates, obtained on small woody macrofossils (about 10 mg each) in two separate cores, provide similar estimates of pre-settlement sedimentation rates in Lake Pepin. The dates, 400 ± 50 ¹⁴C yr. BP (CAMS-39310) and 740 ± 50 ¹⁴C yr. BP (CAMS-39311) occur at sediment depths of 261 cm in core III.2 and 327 cm in core III.1 (Fig. 7). These depths correspond, respectively, to magnetic susceptibility features (mf) 7.5 and 8. Calibration to a calendrical time-scale (Stuiver & Reimer, 1993) yields dates of 1473 (1443-1621; 1 sigma) for mf-7.5 (core III.2) and 1284 (1261-1295; 1 sigma) for mf-8 (core II.1); the larger uncertainty for the younger date results from multiple intercepts of the ¹⁴C calibration curve. Age-depth profiles for each of the two dated cores show fairly constant sediment accumulation rates between calibrated ¹⁴C dates and dated magnetic features 7, 7.5 and 8. Sedimentation rates between ¹⁴C dates and mf-7 (dated to 1830) are 0.19 cm/yr in core II.1 and 0.18 cm/yr in core III.2. These rates are similar to those reported by Wright *et al.* (1998) for a 16 meter long core collected mid-channel between transects II and III. (0.15 cm/yr).

Cesium-137 Dating: Activity profiles for ¹³⁷Cs show well-defined peaks in all 20 analyzed cores corresponding to the 1964 maximum in atmospheric deposition from nuclear testing (Fig. 8). Preliminary counts on several of the detailed cores indicated that the ¹³⁷Cs peak was closely aligned with magnetic feature 2, and only a few counts centered around mf-2 were needed to define the ¹³⁷Cs maximum in subsequent analysis of the top/bottom cores. The core depth of the ¹³⁷Cs marker generally exceeds 1 meter in the upper end of the lake (transect I), which is equivalent to a mean sedimentation rate of more than 3 cm/yr over the last three decades. The sediment column overlying the ¹³⁷Cs peak tends to decrease down river, although there is considerable variation within each transect. In transect V at the lower end of the lake the 1964 ¹³⁷Cs marker occurs between 0.4 and 0.5 meters, which still represents an annual sedimentation rate in excess of 1 cm. These high rates of burial and the high silt/clay content of Pepin sediments tend to limit diffusion (or mixing) of sedimentary ¹³⁷Cs and create near-ideal conditions for preservation of sharply defined ¹³⁷Cs peaks (Van Metre *et al.*, 1997).

Lead-210 Dating: Total ²¹⁰Pb activity shows a roughly monotonic decline with sediment depth in the upper strata of cores from mid- and lower-lake transects (III-V), while activity profiles in the upper lake transects (I and II) are more irregular and exhibit numerous kinks and flat spots (Fig. 9). Surface activities are quite low, especially in upper lake cores, and range from 5.7 pCi/g in core V.1 to 2.0 pCi/g in core I.2. Background (supported) ²¹⁰Pb is well defined at depth in all cores by numerous strata with near constant activity. Supported ²¹⁰Pb values are lowest in cores from the upper end of the lake (0.9 pCi/g in transect I) and increase down lake (1.5 pCi/g in transect V). However, the two dated cores from each transect show virtually the same supported ²¹⁰Pb, providing evidence that supported activities are reasonably constant for any given reach of the lake. The sediment depth to which unsupported ²¹⁰Pb is found is greatest in upper lake cores (> 2 meters in transect I) and shallowest in lower-lake cores (1 meter in transect V), which is consistent with the down-lake decrease in sedimentation rates indicated by other stratigraphic analyses (magnetic susceptibility, pollen, ¹³⁷Cs).

The average lake-wide flux of ²¹⁰Pb to Lake Pepin is 2.3 pCi cm⁻² yr⁻¹, or about 5 times the mean atmospheric deposition rate for the region (0.45 pCi cm⁻² yr⁻¹) (Urban *et al.*, 1990). Although a large portion of the ²¹⁰Pb load to Lake Pepin is clearly delivered by the Mississippi River, the annual lake-wide accumulation (1.7 Ci) represents less than 1% of the ²¹⁰Pb falling on the watershed -- assuming an in-lake retention of 75%, similar to particulate matter (James *et al.*, 1998). It appears that most atmospheric ²¹⁰Pb is retained in the watershed, and what little does make it to Lake Pepin can be largely attributed to direct atmospheric deposition to the surface of the Mississippi River and its tributaries (although some portion of the ²¹⁰Pb is certainly delivered by soil erosion).

All activity profiles show flat spots and other changes in slope that represent historic changes in sediment flux to the core sites (Fig. 9). Such data are best interpreted using the constant rate of supply (c.r.s.) model, which assumes a constant flux of ²¹⁰Pb to the core sites but allows sediment accumulation to vary. The assumption of a constant ²¹⁰Pb input is generally valid for lakes that obtain most of their ²¹⁰Pb by direct atmospheric deposition. However, Lake Pepin probably receives a substantial flux of ²¹⁰Pb as part of the sediment load from the Mississippi River, and the input to individual core sites may have varied over time with changes in river flow or human alterations of its very large catchment. Lead-210 dating of these cores is also uncertain because of the very low values for unsupported ²¹⁰Pb (calculated as the difference between total and supported ²¹⁰Pb). Unsupported activities are particularly low in Pepin sediments, because high sediment fluxes, especially in the upper lake, severely dilute the ²¹⁰Pb inputs. Although the ²¹⁰Pb profiles from Lake Pepin can be interpreted using the c.r.s. model, the resulting dates must be compared with independent dating markers to ascertain their reliability.

Cesium-137 and pollen analysis (together with LOI) provide two dating markers with which the ²¹⁰Pb results may be compared: the 1964 ¹³⁷Cs peak and the 1860 *Ambrosia* rise. In addition, two peaks in magnetic susceptibility (mf-4 and mf-6) are readily identified in all cores and can be used to compare the synchronicity of ²¹⁰Pb dates at depths intermediate between the ¹³⁷Cs and pollen markers. This comparison clearly shows that most ²¹⁰Pb chronologies deviate substantially from one another and from the known dates (Fig. 10). Lead-210 dates at the depth of the ¹³⁷Cs peak are either significantly older or younger than 1964. There is an equally large scatter about the magnetic markers, with mf-4 centered at 1940 and mf-6 around 1890. Lead-210 chronologies from only four cores extend to 1860, and the dates at this maker have very large uncertainty (25-70 years).

The most likely explanation for the discrepancies among core chronologies is a changing flux of ²¹⁰Pb to individual core sites. A number of post-settlement changes to Lake Pepin,

including increased sediment loads, changes in flow regime, basin infilling, and delta progradation, could have altered the pattern of sediment deposition and 210 Pb accumulation within the basin. The c.r.s model can be modified to accommodate changes in core-specific 210 Pb flux by integrating 210 Pb inventories from time-synchronous units of a representative set of cores (Oldfield & Appleby, 1984). The primary assumption here is one of constant 210 Pb input to the entire lake basin rather than to individual core-sites. In the case of Lake Pepin, magnetic susceptibility features mf-2 (the 137 Cs peak), mf-4, mf-6, and the *Ambrosia* rise (or LOI equivalent) provide the depth markers for synchronizing the 10 analyzed cores. In this basin-wide c.r.s. model, $B(\xi)$ is the total unsupported 210 Pb beneath the entire sediment layer defined by depth marker ξ , and the age t of marker ξ is given by:

$$B(\xi) = B(0) \cdot e^{-kt} \tag{1}$$

where k is the decay constant for 210 Pb (0.03114 yr $^{-1}$). $B(\xi)$ and B(o) are calculated by summing the corresponding 210 Pb inventories for individual cores weighted by the portion of the depositional basin each core represents.

The basin-wide c.r.s. model yields the following dates for Lake Pepin: mf-2 (1967), mf-4 (1940), mf-6 (1890), and the *Ambrosia* rise (1841). The ²¹⁰Pb date of 1967 on mf-2 is very close to that determined independently by ¹³⁷Cs (1964), lending support to the basin-wide c.r.s. model. However, the pollen date is clearly too old, indicating that a small amount of unsupported ²¹⁰Pb is missing from the basin-wide inventory. This missing amount, which probably resulted from the difficulty of estimating near-background values of unsupported ²¹⁰Pb in the upper lake cores, can be calculated by fitting the lowermost ²¹⁰Pb date at the *Ambrosia* rise to 1860. The effect of this small addition of unsupported ²¹⁰Pb on the other dates is negligible except for a slight shift of mf-6 to 1894.

Detailed 210 Pb chronologies were established for individual cores by fitting the c.r.s model to fixed dates at the four stratigraphic markers: mf-2 (1964), mf-4 (1940), mf-6 (1895) and the *Ambrosia* rise (1860). Dates between fixed markers ξ_1 and ξ_2 were determined by calculating the unsupported 210 Pb inventory below the lower of the two markers $A(\xi_2)$ as:

$$A(\xi_2) = \frac{A(i)}{\left(e^{k \cdot \Delta t} - 1\right)} \tag{2}$$

where A(i) is the ²¹⁰Pb inventory for the interval between ξ_1 and ξ_2 , and Δt is the time between the two fixed dates. The ²¹⁰Pb inventory above the upper dating marker $A(\xi_1)$ of age t was calculated as:

$$A(\xi_1) = \left[A(i) + A(\xi_2)\right] \cdot \left(e^{kt} - 1\right) \tag{3}$$

and the age at core depth x between ξ_1 and ξ_2 is given by:

$$A(x) + A(\xi_2) = \left[A(\xi_1) + A(i) + A(\xi_2) \right] \cdot e^{-kt}$$
 (4)

where A(x) is the ²¹⁰Pb inventory between x and ξ_2 . The effect of these calculations is to adjust the ²¹⁰Pb inventory for a core at each dated marker under the assumption that the ²¹⁰Pb flux to the core site has remained relatively constant during the interval between markers. Although it is more likely that the ²¹⁰Pb flux varied continuously over time, the resulting dates are closely constrained by the fixed markers and are unlikely to be far from true.

<u>Sediment Accumulation Rates</u>: A composite chronology for each detailed sediment core was compiled from fitted ²¹⁰Pb dates (1860 to present) and ¹⁴C-dated magnetic features for

pre-settlement strata (see above). A date of 1830 was ascribed mf-7 (the settlement horizon), and ¹⁴C dates of 1473 AD and 1284 AD were fixed on mf-7.5 and mf-8, respectively. Sediment flux was calculated as the cumulative dry mass between dated strata divided by the time elapsed. For top/bottom cores a single "modern" sedimentation rate representing the last 35 years was calculated from the ¹³⁷Cs marker, while pre-settlement sediment accumulation was estimated from the dated magnetic features (as in the detailed cores). The pre-settlement sedimentation rate from core IV.4 was applied to two adjacent cores, IV.2 and IV.3, that truncated above mf-7.5; the three cores exhibit parallel susceptibility profiles which suggests similar sedimentation rates in pre-settlement times.

All detailed cores show a dramatic increase in sediment accumulation over the course of the last 200 years (Fig. 11). Accumulation rates double around the time of European settlement, increase gradually during the first few decades of the twentieth century, and then rise sharply between 1940 and 1960. Most of the cores show a clear decline in sediment flux during the 1970s followed by a return to peak values during the last two decades. These trends are broadly synchronous among cores although differing in detail from site to site.

Sediment accumulation rates are highest in cores from the upper end of the lake and decrease progressively down stream. Pre-settlement accumulation rates in cores from transect I range from 1.3 to 2.1 kg m⁻² yr⁻¹, while those from transect V vary from 0.5 to 0.8 kg m⁻² yr⁻¹ (Fig. 11). Modern sedimentation rates are about an order of magnitude higher than pre-settlement rates throughout most of the basin (ranging from 7-15 kg m⁻² yr⁻¹ in transect II to 3-5 kg m⁻² yr⁻¹ in transect V), except in uppermost transect I where present-day rates (20-30 kg m⁻² yr⁻¹) average about 15 times pre-settlement rates. The larger rate increase in cores from transect-I may represent enhanced sedimentation caused by progradation of the Mississippi delta at the upper end of the lake. Higher rates of sediment infilling in the upper reaches of Lake Pepin have been noted by a number of previous investigators (McHenry *et al.*, 1980; Rada *et al.*, 1990; Maurer *et al.*, 1995).

Grain-Size and Sedimentology: Sediments throughout the depositional regions of Lake Pepin are composed primarily of silt-sized clastic minerals (Fig. 12). Below the uppermost 10 cm of core length, sediment density is relatively high (0.4-0.7 g dry sediment cm⁻³) and water content is low (ranging from 70% in transect-I to 50% in transect V). Mean grain size in pre-settlement sediments is decidedly larger in the upper end of the lake (9-11 μ m in transect I) than in mid-lake (7-8 μ m in transect II-IV) or lower lake cores (6-7 μ m in transect V). Grain-size decreases in all cores following settlement, most notably between 1860 and 1910. However, the decrease is most pronounced in the upper lake with the net result that mean grain size becomes a uniform 6-7 μ m throughout the basin by the mid-1900s. A return to coarser grained sediments during the last few decades appears in transects I and II.

Sediment-Phosphorus Content: Total sediment-phosphorus (total-P_{sed}) concentrations are relatively uniform across the depositional basin of Lake Pepin. Pre-settlement values range from 0.6 to 0.8 mg g⁻¹ dry sediment, with lower lake cores (transects IV and V) just slightly higher than mid- and upper lake cores (Fig. 13). Total-P_{sed} shows little variation during presettlement times but increases markedly after about 1850, reaching peak values of 1.6-1.8 mg g⁻¹ between 1950 and 1970. Many of the cores exhibit an earlier peak of 1.0-1.4 mg g⁻¹ between 1890 and 1910, followed by a local minimum around 1920-1930. Total-P_{sed} concentrations decline noticeably after 1970 in upper- and mid-lake cores (transects I-III).

Roughly half of the phosphorus in Lake Pepin sediments is composed of NaOH-extractable P (NaOH-P), with HCl-extractable P (HCl-P) and residual P (organic-P) each making up about one fourth (Fig. 13). NH₄Cl-P averaged less than 5% of total-P_{sed} in all

cores and is summed here with NaOH-P. These proportions vary spatially within the basin as well as stratigraphically within cores. NaOH-P comprises 30-45% of the total in presettlement strata, increases to 60-70% during peak phosphorus concentrations in the 1960s and 70s, and declines toward pre-settlement proportions after 1970. The percentage of HCl-P drops to around 10-15% following settlement, while organic-P shows a variable pattern, rising in some cores (IV.2, III.4, II.1) but not others (V.1, V.4). HCl-P trends toward pre-settlement proportions after 1970 in cores from the upper-lake transects (I and II).

Overall, total-P_{sed} concentrations in recent decades are roughly twice those of pre-settlement times, and much of the increase can be ascribed to labile forms of inorganic-P (NaOH-P). The recent (post 1970) drop in phosphorus content of upper and mid-lake cores corresponds to the increase in sediment grain-size and carbonate content noted earlier. As sedimentation rates also rebound during this time period, it seems likely that P concentrations drop because of dilution from inputs of coarser phosphorus-poor sediments.

Sediment Apportionment: Most of the sediment that settles in Lake Pepin originates from glacial deposits in the Upper Mississippi, Minnesota, and St. Croix river basins (Fig. 1). Because the provenance and lithology of these deposits differ among the three basins, the sediments found in Lake Pepin can be apportioned to their sources based on unique elemental signatures of the mineral portion of the sediment. Sediments originating in the Upper Mississippi River are derived from glacial drift of the Superior Lobe, Rainy Lobe, Wadena Lobe, and St. Louis Sublobe (Des Moines Lobe), whereas sediments from the St. Croix River are composed mainly of Superior Lobe material with a small amount of Grantsburg Sublobe (Des Moines Lobe) material. The Minnesota River contributes sediments almost exclusively of Des Moines Lobe origin.

Because both the Upper Mississippi and St. Croix watersheds drain areas composed mainly of Superior Lobe material, the CMB model was unable to differentiate between most individual contributions from these two river systems. Consequently, the St. Croix and Upper Mississippi rivers were considered to be one source for all but seven samples, effectively reducing the number of distinct contributing sources to two: a Minnesota River source and an Upper Mississippi/St. Croix River source.

Modeling results indicate that the Minnesota River has contributed in excess of 70% of sediments to Lake Pepin since pre-settlement times (Fig. 14). However, most cores show an up-core increase in sediment apportionment to the Minnesota River beginning in the mid-1800s. The cores differ in stratigraphic detail as well as the magnitude of overall change, with upper lake cores showing the smallest shift. In core-transects I and II, the average contribution by the Minnesota river rises 4.5% following European settlement (Fig. 15). In the mid-lake transect III the average post-settlement increase is 8.43%, and in the lower lake (transects IV and V) the increase is 11.71%.

Whole-Lake Fluxes: Fluxes for each sediment fraction were calculated for individual core sites as the product of the sediment accumulation rate and the concentration of the fraction in different strata. In turn, lake-wide fluxes were determined by weighting the flux of each core by the portion of the depositional basin it represents (see methods). In this study we calculated whole-lake fluxes at decadal intervals from the present to 1930, at 20-year intervals from 1930 to 1890 and 30-year intervals from 1890-1830; a single average flux was determined for all analyzed strata older than 1830 (Fig. 16). Fluxes were harmonized in this manner because the dating intervals obtained by ²¹⁰Pb were different for each core; the averaged intervals are consistent with the temporal resolution of geochemical analysis and core dating.

Annual sediment accumulation in Lake Pepin has increased by more than an order of magnitude, from 79,000 t during pre-settlement times to 876,000 t in the 1990s (Fig. 16). The accumulation increase is relatively gradual during the first century following settlement, but accelerates markedly during the 1940s and 50s. Sediment flux appears to plateau at about 700,000 t yr⁻¹ after 1960, with a modest drop during the 1970s and a small peak in the 1990s. The present-day rate is very close to mass-balance estimates of sedimentation (680,000 t yr⁻¹) based on inflow/outflow measurements for 1994-96 (James *et al.*, 1998). Moreover, the sediment core results for 1990-96 integrate high flow years (e.g. 1993) and hence should be expected to be somewhat higher than 1994-96. The convergence of these two estimates of modern mass flux lends considerable weight to historical estimates of sediment accumulation from our retrospective study.

Results from CMB modeling of heavy-mineral elemental signatures indicate that most of the increase in sediment loading to Lake Pepin can be attributed to sources in the Minnesota River basin (Fig. 15). Prior to 1830, whole-basin sediment inputs from the Minnesota River (67,600 t y⁻¹) -- calculated as the product of sediment apportionment and mass accumulation rates for each core -- were four times greater than loads from the Mississippi and St. Croix rivers combined (16,000 t yr⁻¹). Today the Minnesota River contributes 841,000 t yr⁻¹, or more than 7 times the combined load of the other two basins. Thus while the loadings from the combined St. Croix and Upper Mississippi watersheds have increased 7-fold over the last 160 years, the sediment load from the Minnesota watershed has increased more than 12-fold in the same time period.

Fluxes of the major sedimentary components -- organic matter, carbonates, and inorganic matter (clastics) -- parallel the historic changes in total sediment accumulation (Fig. 16). Organic matter increases from a pre-settlement flux of 5,000 t yr⁻¹ to a modern rate of 74,000 t yr⁻¹, while clastic inputs increase from 64,000 to 682,000 t yr⁻¹. Carbonates show a slightly different pattern, holding relatively constant at about 20,000 t yr⁻¹ from settlement to 1940 and then rising dramatically thereafter to peak rates of 119,000 t yr⁻¹ in the 1990s. The rapid increase in carbonate flux since 1940 may reflect increased algal productivity in Lake Pepin, which could drive calcite precipitation though greater uptake of CO₂ (Engstrom & Swain, 1986; Hodell *et al.*, 1998). The carbonate flux closely tracks whole-lake phosphorus trends and diatom-inferred TP for the water column (see below), suggesting that the most dramatic trophic changes in Lake Pepin have occurred since c. 1945.

The lake-wide flux of total-P_{sed} mirrors the historic increase in total sediment accumulation (Fig. 16). Modern rates (800-900 t yr⁻¹) are about 15 times pre-settlement rates (60 t yr⁻¹) and are relatively constant following a 1960s peak of 1000 t yr⁻¹. All of the various P fractions show the same broad post-settlement increase, although differing somewhat in detail. NaOH-P, the largest fraction, exhibits the greatest change, rising to a peak flux of 680 t yr⁻¹ during the 1960s and declining to present-day rates of 480 t yr⁻¹; the 1960s flux is more than 25 times the pre-settlement accumulation rate (25 t yr⁻¹). The flux profile for HCl-P is very similar to that for carbonate, holding steady at 40-50 t yr⁻¹ from 1830 to 1940 and then rising sharply to 160 t yr⁻¹ during the next two decades; the 1990s carbonate peak also shows up in HCl-P accumulation. The distinctive features of the HCl-P profile and their correspondence with carbonate flux are a clear indication that the HCl-P fraction is closely bound to sedimentary carbonates. The annual flux of residual (organic) P increases about 20 times from pre-settlement (10 t) to present-day values (200 t); fluxes have been relatively constant since about 1950.

Our estimate of present-day sedimentation rates for total- P_{sed} (920 t yr⁻¹) is somewhat higher than mass-balance calculations would indicate. Annual retention of phosphorus in Lake Pepin sediments averaged 510 t from 1994 to 1996 based on the difference between total-P inflow (3930 t yr⁻¹) and outflow (3420 t yr⁻¹) (James *et al.*, 1998). However, these

monitoring data do not include the 1993 flood year in which exceptional sediment loading almost certainly raised the average total-P_{sed} flux to the stratigraphic record for 1990-96. Moreover, the mean in-lake retention estimated by James *et al.* (1998) is only 13% of annual inputs, meaning that small uncertainties in estimating inflows or outflows could also explain some of the difference between their estimate of P accumulation and ours. These differences notwithstanding, the convergence of two totally independent estimates of modern P sedimentation provides compelling evidence that the stratigraphic record offers a reliable measure of historical phosphorus loading to the upper Mississippi River.

Diatom Stratigraphy and Inferred Lakewater-TP: A total of 145 taxonomic entities were identified in the 10 stratigraphic samples analyzed for fossil diatoms. Sample intervals were 4-cm thick and integrated from 2 to 12 years of accumulation. Samples older than 1920 were very low in diatoms with the lowest numbers occurring around 1900. Diatom microfossils were slightly more numerous and most heavily silicified in the bottom sample (c. 1760). Preservation of the more lightly silicified taxa did not appear to be a problem, as intact valves of *Asterionella formosa* were present to a core depth of 92-96 cm (an uncounted level dating to c. 1880). The good preservation found in core V.1 stands in contrast to that reported previously by Ohl (1992) for a 1989 core from lower Lake Pepin. Ohl found numerous diatom fragments and low abundance at depth in his profile, which he attributed to poor preservation.

Actinocyclus normanii f. subsalsa was identified at 64-68 cm (c. 1940) and in several samples above that level. Aulacoseira species are present throughout, but are most abundant in samples dated between 1920 and 1964 (Fig. 17). With the exception of one occurrence in the 1760 sample, Cyclotella meneghiniana is present around 1900 and increases in abundance upwards in the core. Benthic Fragilaria species are most abundant in the lower samples and decline to $\leq 2\%$ after 1920. Stephanodiscus species begin to increase after 1900 and are most abundant in the upper three samples (1975 to the present).

Benthic diatom species have their highest abundance in the two pre-settlement samples (Fig. 17). At 1860, benthic and planktonic species each compose 50% of the diatom assemblage. Planktonic diatoms continue to increase above this level and constitute ≥90% of the assemblage from 1940 to the present. Chrysophyte cysts comprise ca. 10% of the microfossil assemblage in the lower samples, reaching a maximum of 16% around 1900. Their abundance declines rapidly after 1920. Most of the diatoms found in core V.1 are characteristic of lentic environments and are thought to derive principally from in-lake production.

A total of 30 taxa from the Lake Pepin data were used for lake-water total-phosphorus (TP) reconstruction after matching species with published optima from the Bennion (1994) calibration set; these taxa represented 55-73.5 % of the total diatoms in each sample. After addition of species optima from Reavie *et al.* (1995), the number of taxa used in the reconstruction rose to 36 and represented 77-90% of the total diatom abundance (Fig. 17). In particular, *Aulacoseira subarctica* (O. Müller) Haworth, is not present in the Bennion data set, but is included in those of Reavie *et al.* (1995). It is an important taxon from 1900 to 1940 and constitutes 13-27% of the total diatom assemblage in these samples.

Two different TP reconstructions were produced from the Pepin diatom data. The first (TP-1) used additional species optima from Reavie *et al.* (1995) only for those taxa absent from the Bennion data set. The second TP inference (TP-2) was constructed using optima from Reavie *et al.* for all *Aulacoseira* and *Fragilaria* species (Fig. 17). These taxa are infrequent among the high-TP lakes sampled by Bennion, but are well represented at the lower TP range of the Reavie *et al.* data set. Because these taxa are also dominant in the deeper strata of the Pepin core, the second TP inference (TP-2) is probably a more accurate

reconstruction of pre-settlement nutrient conditions than TP-1. TP values from this second inference model are applied to pre-1860 strata in the following discussion and are used for reconstructing a phosphorus mass-balance for the lake (see below).

Diatom-inferred TP values are lowest (49-55 μ g l⁻¹, TP-2) in the bottom two samples from 1760 and 1800 (Fig. 17). Mean TP concentrations increase immediately following settlement, reaching a peak of 125 μ g l⁻¹ (TP-1) around 1900, and then decline to values of 111-121 μ g l⁻¹ between 1920 and 1940. A sharp rise in mean TP to 173 μ g l⁻¹ occurs between 1940 and 1975, followed by a continuous increase to the present (205 μ g l⁻¹). The modern value for diatom-inferred TP is virtually identical to measured summer TP concentrations in the outflow of Lake Pepin for the period 1977-1991 (213 ± 23 μ g l⁻¹; χ ± s.e.) (Heiskary & Vavricka, 1993).

The high percentage of benthic diatom species and chrysophyte cysts in the lower samples of the core indicate that prior to European settlement, clear water conditions existed for at least part of the year in Lake Pepin. This type of flora is commonly found in post-glacial sediments (Haworth, 1972; Bradbury & Whiteside, 1980) and low-nutrient Arctic lakes (Smol, 1993). The *Aulacoseira* species present are indicative of moderate nutrient levels and a turbulent environment. These taxa form large chains and are dependent on turbulent conditions such as those found in a large river system, to maintain themselves in the photic zone (Bradbury, 1975). The heavy silicification of diatom valves in these early sediments indicates that silica (SiO₂) concentrations were not limiting and that Si:P ratios were high (Stoermer *et al.*, 1985; Stoermer *et al.*, 1989). Mean TP values are lowest in these samples (49-55 μ g l⁻¹; TP-2).

Reconstructed mean TP concentrations increase to 83 μ g l⁻¹ (TP-2) by 1860 and reflect the influence of European settlement and land clearance in the basin. Diatom microfossil numbers are low in samples just prior to and immediately after this level. The low numbers indicate probable dilution caused by higher sediment loads. A peak in inferred TP occurs around the turn of the century (125 μ g l⁻¹) and corresponds with the first increases in *Stephanodiscus* species and the beginning rise in *Aulacoseira* abundance. These diatom groups have high and medium-high phosphorus optima, respectively. *Cyclotella meneghiniana* also occurs at this level and increases to the surface of the core. This taxon is associated with slightly alkaline environments (Fritz *et al.*, 1993) and has been reported from nearshore areas in nutrient-enriched Lakes Erie and Ontario (Stoermer, 1978). Its occurrence in Lake Pepin indicates an increase in nutrient loading and possibly salinity.

The onset of European settlement and the rise in mean TP concentrations corresponds with increased abundance of planktonic taxa in the core (Fig. 17). Higher abundance of planktonic forms increases shading, and coupled with the increased turbidity from watershed erosion, results in a loss of clear-water benthic diatom taxa and chrysophytes (after 1900). Jeppesen *et al.* (1990) suggested that crossing a critical threshold level of 80-150 μ g l⁻¹ TP results in light limitation and loss of submerged macrophytes in shallow lake systems. This same mechanism may be responsible for the loss of benthic diatom populations in Lake Pepin, although increased turbidity from suspended solids is equally likely. Smol (1985) showed that loss of chrysophyte cysts commonly occurs in sediments with increased anthropogenic activity. The decline in inferred-TP around 1920 (111 μ g l⁻¹) corresponds with a local minimum in sedimentary P concentrations (see above) and may indicate reduced nutrient loading resulting from a temporary cessation of watershed disturbance and possibly from reduced flows in the Mississippi during the dust-bowl era (see below).

The rise in mean TP after 1965 (from 142 to 173 μ g l⁻¹) signals another major perturbation in the river. Diatom assemblages shift from an *Aulacoseira* (medium-high optima)

dominated system to a *Stephanodiscus* (high optima) dominated one. The rise in *Stephanodiscus* species indicates increased nutrient loading to the lake. Additionally, *Actinocyclus normanii* f *subsalsa* first occurs at the 1940 level. It is a taxon common in highly polluted nearshore regions and bays in the Great Lakes and is indicative of areas with high industrial activity (Stoermer, 1978). Mean TP concentrations increase continuously from 1975 to the present (from 173 to 204 μ g l⁻¹). The uppermost three samples are dominated by small *Stephanodiscus* species with high TP optima that are common in highly eutrophic systems. No apparent decrease in TP concentrations occurs at the surface, indicating that nutrient levels have continued to rise. However, historical data on summer TP concentrations in Lake Pepin (1976-1991) do not show a systematic increase, but fluctuate about a mean of 240 ppb (Heiskary & Vavricka, 1993); a mean of 180 μ g l⁻¹ was recorded in 1964-65 (FWPCA, 1996). This discrepancy with diatom-inferred TP may simply reflect the coarse temporal resolution of the diatom data, or may indicate that the diatoms are responding to other factors (e.g. light limitation caused by suspended sediment) in addition to mean summer phosphorus concentrations.

Phosphorus Mass-balance: Because of its relatively short residence time (6-47 days), Lake Pepin retains in its sediments a relatively small portion of the total phosphorus load from the Mississippi River. Current estimates from inflow/outflow measurements for 1994-96 place in-lake retention at about 13%, with the remaining 87% lost though outflow (James *et al.*, 1998). If we assume that P retention has remained relatively constant over time, then historical phosphorus loading -- estimated by dividing annual P accumulation in Lake Pepin sediments by 13% -- has increased progressively from 430 t yr⁻¹ in pre-settlement times to 7000 t yr⁻¹ at the present day. However, these estimates are likely to be in error, because relatively small changes in retention could have a large effect on the rate of P sedimentation. Change in discharge (which would affect residence time), eutrophication and associated changes in lake-bottom redox conditions (which could enhance P-release from sediments), and increased sedimentation rates (which might have the opposite effect and enhance P burial) have occurred since European settlement and have likely altered P retention in Lake Pepin.

A preferable alternative would be to estimate historical losses of phosphorus from Lake Pepin through Mississippi River outflow and calculate P loading to the lake as the sum of sedimentation and outflow. Outflow losses can be calculated from historical flow records and reconstructions of water-column phosphorus concentrations from fossil diatoms in Pepin sediments. Such methods, applied by Rippey and Anderson (1996) to a small eutrophic lake in Northern Ireland, show reduced P sedimentation during periods of increased lake-water TP, illustrating that assumptions of constant P-retention can be highly misleading.

Discharge records for the Mississippi River at the USGS gauging station at Prescott, WI (below the confluence with the St. Croix River) extend back to 1928. Earlier flows at Prescott can be estimated from those at St. Paul, MN (back to 1892) and St. Croix Falls, WI (back to 1910 and intermittently to 1902) by scaling the discharge at those upstream stations to flows at Prescott for the period of common record (1928-1996). This composite record (Fig. 18) shows large variations in mean annual flow, with persistently low values during the 1920s and 1930s and substantially higher flows from the 1940s onward, especially during the last three decades. Flows prior to 1892 must be extrapolated from the period of record; our best estimate is the average discharge from 1892 to 1930, which excludes the extreme low flows of the Dustbowl era and high flows of the late 20th century that resulted, in part, from hydrological alterations and land-use changes (wetland destruction, ditching, and drainage tiles).

Reconstructed phosphorus outflow from Lake Pepin, calculated at decadal intervals as the product of diatom-inferred TP and mean discharge, shows a gradual rise from 700 t yr⁻¹ in pre-settlement times to 2200 t yr⁻¹ around the turn of the century (Fig. 19). Phosphorus outflow drops sharply during the low-flow decades of the 1920s and 30s, returns to 1900-2300 t yr⁻¹ between 1940 and 1970, and then rises sharply to present-day values of 4500 t yr⁻¹. Our estimate for the 1990s compares well with direct measurement of P outflow (3400 t yr⁻¹) for 1994-96 (James *et al.*, 1998).

Historical phosphorus loading to Lake Pepin from the Mississippi River, the sum of P outflow and P sedimentation, increases by a factor of three from 1830 (800 t yr⁻¹) to 1900 (2400 t yr⁻¹), declines during the 1920s and 30s to a minimum of ca. 1200 t yr⁻¹, and then rises continuously from the 1940s (2500 t yr⁻¹) to the present decade (5450 t yr⁻¹) (Fig. 19). Again, our 1990s estimate for P loading (which includes the 1993 flood year) is only slightly higher than P inflow measurements by James *et al.* (1998) for 1994-96 (3900 t yr⁻¹). According to our reconstructions, P loading to Lake Pepin has increased about 7-fold from 1830 to the present. Changes in discharge account for some of the trend, and certainly explain much of the decade-to-decade variation, but the overall increase is driven by rising phosphorus concentrations in the water, and ultimately anthropogenic impacts in the watershed.

These mass-balance calculations show that in-lake retention of total-P has not remained constant over time, but has risen markedly since the turn of the century (Fig. 19). Prior to European settlement, only 7% of the P inflow was sedimented in Lake Pepin; this value changed little until about 1910 when it rose to 16%. Phosphorus retention peaked at 23 to 32% between 1930 and 1970 and then declined progressively to present-day values of 17%. Some of this change appears to be driven by discharge, particularly the high retention rates of the 1930s (30%) and the recent decline of the last three decades. Longer residence times during low flows (e.g. 1930s) should favor higher retention rates, and conversely rapid flushing during high flows (1970s-1990s) should retard phosphorus sedimentation. However, differences in flow alone cannot explain the overall increase in P retention in the 20th century, and other factors favoring P sedimentation must be invoked. The most likely explanation is the large increase in sediment accumulation that begins in the early 1900s. Increasing sedimentation rates should enhance P retention either through a higher particle flux (scavenging P from the water column) or more rapid burial (slowing diffusive or advective release of sedimentary P). Eutrophication, which often leads to greater P release through hypolimnetic anoxia (Marsden, 1989), appears not to have affected P retention to nearly the same degree as sediment burial rates or river flows. Although experimental studies show much higher rates of P-release from Pepin sediments under anoxic conditions (MWCC, 1993c), the water column of Lake Pepin is usually oxygenated year-round (Heiskary & Vavricka, 1993), and bottom-water anoxia is restricted to low-flow conditions.

<u>Lake Infilling</u>: Sediment from the Mississippi River has been filling Lake Pepin since the lake was formed 10,000 ¹⁴C years ago. The lake is thought to have originally extended upstream as far as St. Paul according to bridge-boring records (Zumberge, 1952), and a 16-m core of lacustrine sediment recovered mid-lake by Wright *et al.* (1998) indicates a once much deeper basin. However, the rate of infilling has clearly accelerated in the 170 years since European settlement, greatly shortening the projected life of the lake. Volumetric sediment accumulation, calculated by dividing the dry-mass sedimentation rate by the mean bulk density for each core, indicates that lake-wide infilling has increased from 150,000 m³ yr⁻¹ in pre-settlement times to more than 1,600,000 m³ yr⁻¹ (about 1.6 cm yr⁻¹ over the entire basin) during the 1990s (Fig. 20). This calculation corrects for compaction and de-watering that occurs as sediments are buried by subsequent deposition.

Extrapolating present-day accumulation rates into the future, the remaining volume of Lake Pepin (553 x 10⁶ m³ in 1990) will be filled completely in about 340 years. And given its much higher accumulation rates (3.0 cm yr⁻¹), the shallow upper one-third of the lake (regions I and II, Fig 1) will cease to be useful for recreation or commerce within about a century. It may take somewhat longer than projected to fill the last remnants of the lake, as sediment trapping efficiency will likely decline with loss of volume (Brune, 1953). This obliteration of Lake Pepin will occur in less than one-tenth the time required under presettlement conditions.

Projecting sedimentation rates back in time, lake volume at the onset of European settlement (1830) was 653 x 10⁶ m³, and by 1890 it was reduced to 634 x 10⁶ m³ (Fig. 20). Our volume estimate for 1890 is 52 x 10⁶ m³ smaller than that calculated by Maurer *et al.* (1995) from 1897 bathymetric maps of the Mississippi River (686 x 10⁶ m³). This volumetric difference is equivalent to a mean depth of 0.5 m; that is, the average depth of Lake Pepin according to Maurer *et al.* (1995) was 0.5 m deeper than our sediment cores indicate. This difference is far larger than can be explained by sediment-dating uncertainty, for even if we consider lake volume at the *Ambrosia* rise (1860), the lake was still 41 x 10⁶ m³ smaller than Maurer *et al.* (1995) estimate. It seems far more likely that the error lies with early depth surveys of the Mississippi River rather than the explicit measurements of lake infilling that sediment cores provide. This discrepancy may also explain how McHenry *et al.* (1980) found recent sedimentation rates in Lake Pepin (determined by ¹³⁷Cs) lower than those in the first half of the 20th century (as estimated from an 1895 bathymetric survey).

The rate at which Lake Pepin is filling with sediment has increased by more than an order of magnitude since pre-settlement times. Today about 2.5% of the 1830 lake-volume is lost every decade, while just prior to European settlement the decadal loss was 0.23% (Fig. 20). The cumulative change since 1830 has been a decrease of 17%. Under pre-settlement rates the cumulative loss from 1830 to 1996 would have been just 4%, and it would be another 500 years from now before the lake shallowed to its present-day volume.

SUMMARY AND CONCLUSIONS

<u>Historic Trends</u>: The sediments of Lake Pepin record an order-of-magnitude increase in the loading of suspended solids from the Mississippi River beginning with European settlement c. 1830 and continuing to the present day. Most of this increase can be traced to the Minnesota River basin, which today contributes over 90% of the sediment to Lake Pepin. Modern fluxes are close to one million metric tons annually, and are more than 80% detrital mineral matter. Little of this clastic sediment probably leaves Lake Pepin by outflow, although mass-balance calculations are inconclusive on this matter, as biogenic silica (diatoms) and endogenic carbonates that are produced in the lake are not distinguished from detrital minerals in input/output monitoring. Current estimates of 75% retention of suspended solids (James *et al.*, 1998) provide a minimum estimate of trapping efficiency for clastic materials in the lake.

These sediment loads are currently filling Lake Pepin at an extremely rapid rate. About 17% of the lake volume in 1830 has been replaced by sediment, and if current rates continue, the lake will be filled completely in another 340 years. The upper one-third of the lake, with its higher sediment flux, will be gone in a century. Without this acceleration of sediment loading, Lake Pepin would be on average one meter deeper today and could persist another 4000 years.

Sediment P accumulation in Lake Pepin has increased even faster than the total-sediment flux, and current rates of 900 metric tons annually are more than 15 times pre-settlement rates. Part of this rise represents a seven-fold increase in total-P loading from the

Mississippi River and part reflects more efficient retention of phosphorus inflows by bottom sediments. In the 1990s an average of 5400 metric tons of phosphorus entered Lake Pepin annually, and 17% of this load went to the bottom. Retention rates exceeded 20% from 1930 to 1970, while prior to 1910 only 10% of the external phosphorus load was retained. More efficient trapping of phosphorus in Lake Pepin over the last century probably resulted from higher rates of sediment burial. This observation runs counter to expectations that cultural eutrophication, especially if accompanied by hypolimnetic anoxia, should enhance internal P loading and decrease P-retention (Anderson, 1997). However, thermal stratification is weak in Lake Pepin and bottom-waters are seldom if ever anoxic (Heiskary & Vavricka, 1993). *In situ* pore-water gradients indicate P release from Pepin sediments, but at rates similar to those measured in laboratory experiments under oxic conditions (MWCC, 1993c). Even if present-day release rates are higher than those in the distant past, they clearly have been overwhelmed by the historic increase in gross P sedimentation.

The historic increase in external phosphorus loads has raised total-P levels in waters of Lake Pepin at least four-fold (from 50 to 200 μ g l⁻¹). Increased nutrient availability is recorded by a marked shift in diatom flora from one dominated by clear-water benthic species and mesotrophic planktonic taxa to exclusively planktonic assemblages characteristic of highly eutrophic conditions. Rising planktonic productivity along with increased turbidity from sediment loading ultimately shaded out the benthic algal communities that were more abundant in pre-settlement times.

The most dramatic changes in nutrient and sediment loads to Lake Pepin have occurred since 1940, although gradual increases began shortly following European settlement (c. 1830). Sediment accumulation rates rose sharply between 1940 and 1970 and then leveled off, while phosphorus inflows record their largest increases after 1970. The post-1970 rise in phosphorus loading reflects higher mean flows in the Mississippi River as well as increasing phosphorus concentrations in its waters during the last three decades. The recent increase in total-P inferred from our sediment study is consistent with a rising trend in flow-adjusted summer total-P measured at the inflow to Lake Pepin from 1977 to 1991 (Heiskary & Vavricka, 1993).

<u>Predictions</u>: In addition to providing a baseline for natural (pre-European settlement) conditions in the Mississippi River, results from this study can help predict responses of the system to future reductions in sediment and nutrient loads in a restoration scenario. For example, a key finding from this investigation is that severe eutrophication of Lake Pepin is a relatively recent development and that total-P levels in the lake were about two-thirds their current value prior to 1960. Phosphorus reductions approaching 1950s values might be a reasonable target for remediation efforts and should, using diatoms as indicators, return algal communities to a more desirable composition.

Another question regarding reductions in external nutrient loads is the extent to which the large reservoir of labile P in Lake Pepin sediments might continue to drive algal productivity through internal P loading. Results from this study indicate that this "memory" effect should be relatively short-lived, as diffusive or advective P release would be limited by rapid burial of older sediments. Diffusive P fluxes from the bottom of Lake Pepin appear driven by concentration gradients in the uppermost 2 cm of sediment (MWCC, 1993c), a thickness equivalent to about one year's worth of new accumulation. Sediment mixing, which might also enhance P release, does not occur to any great depth in Pepin sediments (probably less than 2 cm), as shown by sharp stratigraphic detail in magnetic susceptibility, loss-onignition, and ¹³⁷Cs profiles as well as the extremely low densities of burrowing organisms (Chironomid larvae) found in the cores. Hence, a more accurate view of internal P loading

is one of recycling of recent external inputs, rather than a long-term leakage from older Prich strata.

<u>Uncertainties</u>: The task of unraveling the sedimentary record of this large riverine basin was not without difficulties. Sediment dating was a particular problem given the extremely high sedimentation rates that severely dilute the atmospheric flux of ²¹⁰Pb and limit the age to which dates may be reliably extended. Moreover, the flux of ²¹⁰Pb to different parts of the basin appears to have shifted over time, requiring these radiometric data to be integrated stratigraphically throughout the basin. A full array of dating tools including magnetic susceptibility, loss-on-ignition, pollen analysis, ¹³⁷Cs, and ¹⁴C, was ultimately needed to synthesize a reliable sediment chronology. Uncertainties remain, especially for older strata, but the various dating markers constrain one another as do the chronologies for individual cores, and the final results appear robust.

Other difficulties included the strong longitudinal gradient in sediment accumulation that required a large number of cores to characterize. The density of cores employed in this study appears to have been sufficient, for the resulting sediment fluxes for the current decade match closely sediment loads determined independently by input/output monitoring. This success can be partially attributed to the unambiguous delineation of the depositional region of the lake, which is defined by a sharp morphometric scarp separating erosional surfaces from those conformably accumulating fine-grained sediment.

Uncertainties also persist in the phosphorus mass balance. Diatom reconstructions of lakewater TP concentrations are consistent with measured values for the current decade, but past estimates are less secure. There are a number of problems with these numerical methods (Anderson, 1997), not the least of which is the lack of a single regional phosphorus model that adequately encompasses the range of diatom assemblages found in the Lake Pepin record. A synthesis of TP optima from several training sets was used to model the Lake Pepin data, and results, especially for the older strata, are sensitive to the optima of a few key taxa. Likewise, discharge records for the Mississippi River at Lake Pepin do not extend to pre-settlement times, and extrapolating mean flows to these older strata could result in substantial error. Nonetheless, outflow estimates for total-P (the product of diatom-inferred TP and discharge) are proportional to P sedimentation from pre-settlement times to the early 1900s, a period when P retention in Lake Pepin might be expected to be reasonably constant.

Results from this study demonstrate that it is possible to obtain quantitative historical accounts of sediment and nutrient loading in riverine systems from sediment accumulation in impoundments along their courses. Few previous investigations have attempted such a mass-balance approach (Evans & Rigler, 1980; Moss, 1980; Rippey & Anderson, 1996), and none has been done on a major watershed the size of the upper Mississippi River. And because most river impoundments are man-made, the historical records they provide are of generally short duration and do not include pre-impact conditions (Bradbury & Van Metre, 1997; Van Metre *et al.*, 1997). Few major rivers possess natural impoundments like Lake Pepin where sediments have accumulated conformably for thousands of years.

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REFERENCES

- Anderson, N. J., 1997. Reconstructing historical phosphorus concentrations in rural lakes using diatom models. In H. Tunney, O. T. Carton, P. C. Brookes & A. E. Johnston (ed.), Phosphorus Loss from Soil to Water. CAB International, 95-118.
- Anderson, N. J. & B. Rippey, 1994. Monitoring lake recovery from point-source eutrophication: the use of diatom-inferred epilimnetic total phosphorus and sediment chemistry. Freshwat. Biol. 32: 625-639.
- Appleby, P. G. & F. Oldfield, 1978. The calculation of lead-210 dates assuming a constant rate of supply of unsupported ²¹⁰Pb to the sediment. Catena 5: 1-8.
- Bennion, H., 1994. A diatom-phosphorus transfer function for shallow, eutrophic ponds in southeast England. Hydrobiologia 275/276: 391-410.
- Binford, M. W., 1990. Calculation and uncertainty analysis of ²¹⁰Pb dates for PIRLA project lake sediment cores. J. Paleolimnology 3: 253-267.
- Birks, H. J. B., S. Juggins & J. M. Line, 1990. Lake surface-water chemistry reconstruction from paleolimnological data. In B. J. Mason (ed.), The Surface Water Acidification Program. Cambridge University Press, Cambridge, UK: 313.
- Blomqvist, S., 1985. Reliability of core sampling of soft bottom sediment—an in situ study. Sedimentology 32: 605-612.
- Blomqvist, S., 1991. Quantitative sampling of soft-bottom sediments: problems and solutions. Mar. Ecol. Prog. Ser. 72: 295-304.
- Bradbury, J. P., 1975. Diatom stratigraphy and human settlement in Minnesota. Geol. Soc. Am. Spec. Pap. 171: 1-74.

- Bradbury, J. P. & P. C. Van Metre, 1997. A land-use and water-quality history of White Rock Lake reservoir, Dallas, Texas, based on paleolimnological analyses. J. Paleolimnology 17: 227-237.
- Bradbury, J. P. & M. W. Whiteside, 1980. Paleolimnology of two lakes in the Klutlan Glacier region, Yukon Territory, Canada. Quat. Res. 14: 149-168.
- Brezonik, P. L. & D. R. Engstrom, 1998. Modern and historic accumulation rates of phosphorus in Lake Okeechobee, Florida. J. Paleolimnology 20: 31-46.
- Brune, G. M., 1953. Trap efficiency of reservoirs. Trans. Am. Geophys. Union 34: 407-418.
- Dean, W. E., Jr., 1974. Determination of carbonate and organic matter in calcareous sediments and sedimentary rocks by loss on ignition: comparison with other methods. J. Sed. Petrol. 44: 242-248.
- Dearing, J. A., J. K. Elner & C. M. Happey-Wood, 1981. Recent sediment flux and erosional processes in a Welsh upland lake-catchment based on magnetic susceptibility measurements. Quat. Res. 16: 356-372.
- Eakins, J. D. & R. T. Morrison, 1978. A new procedure for the determination of lead-210 in lake and marine sediments. Int. J. Appl. Radiat. Isot. 29: 531-536.
- Engstrom, D. R. & E. B. Swain, 1986. The chemistry of lake sediments in time and space. Hydrobiologia 143: 37-44.
- Engstrom, D. R. & H. E. Wright, Jr., 1984. Chemical stratigraphy of lake sediments as a record of environmental change. In E. Y. Haworth & J. W. G. Lund (ed.), Lake Sediments and Environmental History. Univ. Minnesota Press, Minneapolis: 11-67.
- Evans, R. D. & F. H. Rigler, 1980. Measurement of whole lake sediment accumulation and phosphorus retention using lead-210 dating. Can. J. Fish. Aquat. Sci. 37: 817-822.
- Faegri, K. & J. Iversen, 1975. Textbook of Pollen Analysis. Hafner Press, New York, 295 pp.
- Fritz, S. C., S. Juggins & R. W. Battarbee, 1993. Diatom assemblages and ionic characterization of lakes of the northern Great Plains, North America: a tool for reconstructing past salinity and climate fluctuations. Can. J. Fish. Aquat. Sci. 50: 1844-1856.
- FWPCA, 1996. Report on Water Quality Investigations of the Mississippi, Minnesota, and St. Croix Rivers. Federal Water Pollution Control Administration, Great Lakes Region.
- Grimm, E. C., 1984. Fire and other factors controlling the big woods vegetation of Minnesota in the Mid-nineteenth century. Ecol. Monogr. 54: 291-311.
- Hall, R. I. & J. P. Smol, 1992. A weighted-averaging regression and calibration model for inferring total phosphorus concentration from diatoms in British Columbia (Canada) lakes. Freshwat. Biol. 27: 417-434.

- Haworth, E. Y., 1972. Diatom succession in a core from Pickerel Lake, northeastern South Dakota. Geol. Soc. Am. Bull. 83: 157-172.
- Heiskary, S. & M. Vavricka, 1993. Lake Pepin Water Quality. Section 3, Mississippi River Phosphorus Study. Minnesota Pollution Control Agency, St. Paul, MN, 105 pp.
- Hieltjes, A. H. M. & L. Lijklema, 1980. Fractionation of inorganic phosphates in calcareous sediments. J. Environ. Qual. 9: 405-407.
- Hodell, D. A., C. L. Shelske, G. L. Fahnenstiel & R. L.L., 1998. Biologically induced calcite and its isotopic composition in Lake Ontario. Limnol. Oceanogr. 43: 187-199.
- James, W. F., J. W. Barko & H. L. Eakin, 1998. Synthesis of Phosphorus/Seston Fluxes and Phytoplankton Dynamics in Lake Pepin (Upper Mississippi River), 1994-1996. U.S. Army Engineer Waterways Experiment Station, Spring Valley, WI.
- Jeppesen, E., M. Søndergaard, E. Mortensen, P. Kristensen, B. Riemann, H. J. Jensen, J. P. Müller, O. Sortkjaer, J. P. Jensen, K. Christoffersen, S. Bosselmann & E. Dall, 1990.
 Fish manipulation as a lake restoration tool in shallow, eutrophic temperate lakes 2: threshold levels, long-term stability and conclusions. Hydrobiologia 200/201: 219-227.
- Justic, D., N. N. Rabalais & R. E. Turner, 1995. Stoichiometric nutrient balance and origin of coastal eutrophication. Mar. Pollution Bull. 30: 41-46.
- Marsden, M. W., 1989. Lake restoration by reducing external phosphorus loading: the influence of sediment phosphorus release. Freshwat. Biol. 21: 139-162.
- Maurer, W. R., T. O. Claflin, R. G. Rada & J. T. Rogala, 1995. Volume loss and mass balance for selected physicochemical constituents in Lake Pepin, upper Mississippi River, USA. Reg. Rivers 11: 175-184.
- McHenry, J. R., J. C. Ritchie & C. M. Cooper, 1980. Rates of recent sedimentation in Lake Pepin. Water Resour. Bull. 16: 1049-1056.
- Moss, B., 1980. Further studies on the palaeolimnology and changes in the phosphorus budget of Barton Broad, Norfolk. Freshwat. Biol. 10: 261-279.
- MWCC, 1993a. Mississippi River Phosphorus Study, Sections 1-6. Metropolitan Waste Control Commission, St. Paul, MN.
- MWCC, 1993b. Point and Nonpoint Source Phosphorus Contributions. Section 4, Mississippi River Phosphorus Study. Metropolitan Waste Control Commission, St. Paul, MN, 66 pp.
- MWCC, 1993c. River and Lake Sediment. Section 5, Mississippi River Phosphorus Study. Metropolitan Waste Control Commission, St. Paul, MN, 130 pp.
- MWCC, 1993d. Water Quality of the Mississippi River and Tributaries, 1976-1991. Section 3, Mississippi River Phosphorus Study. Metropolitan Waste Control Commission, St. Paul, MN, 136 pp.
- MWCC, 1994. Water Quality Analysis of the Lower Minnesota River and Selected Tributaries: River (1976-1991) and Nonpoint Source (1989-1992) Monitoring.

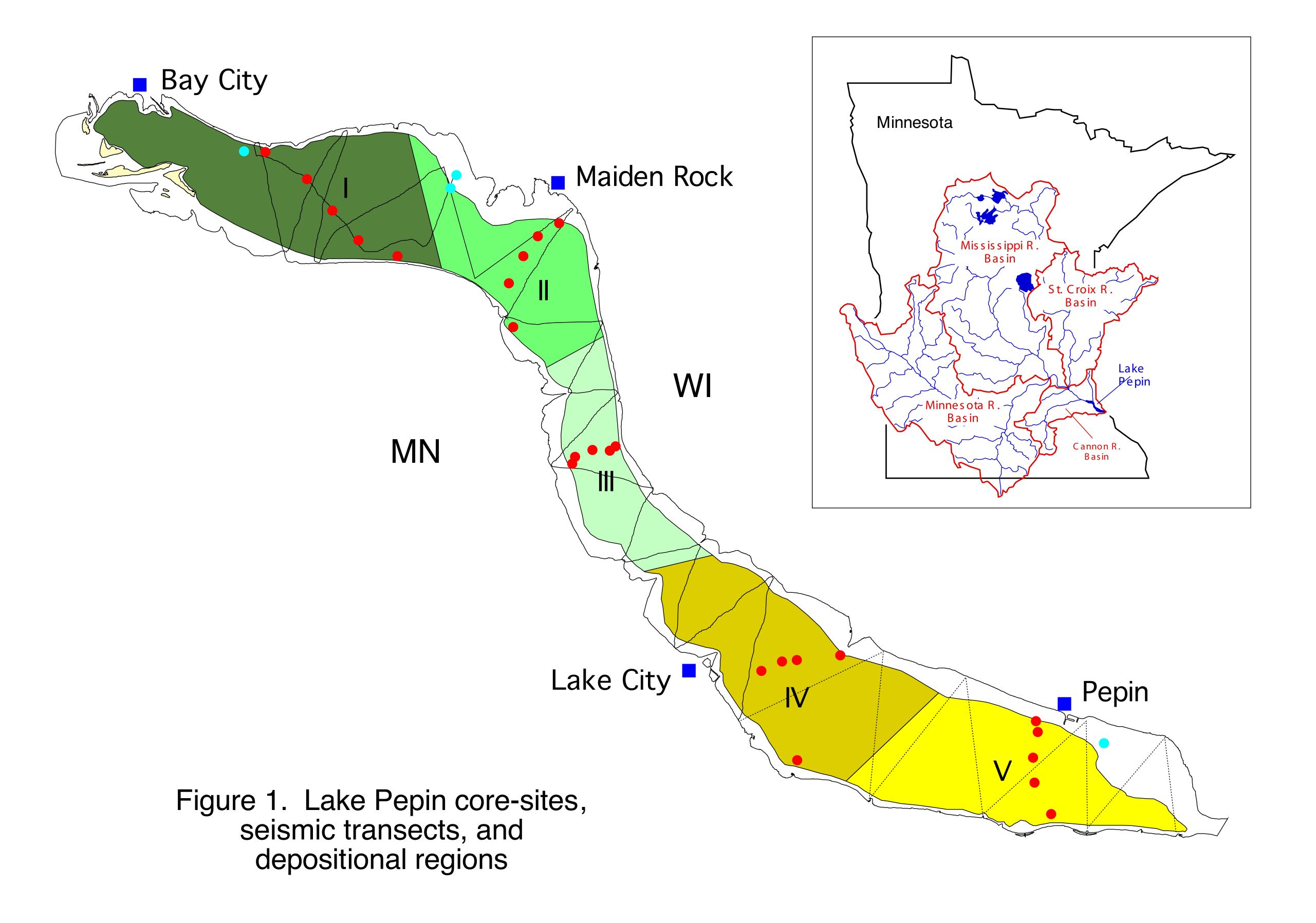
- Volume 1, Report QC-93-267. Metropolitan Waste Control Commission, St. Paul, MN, 142 pp.
- Ohl, L., 1992. Historical Trends in the Trophic State of Lake Pepin. Metropolitan Waste Control Commission, St. Paul, MN.
- Oldfield, F. & P. G. Appleby, 1984. Empirical testing of ²¹⁰Pb-dating models for lake sediments. In E. Y. Haworth & J. W. G. Lund (ed.), Lake Sediments and Environmental History. University of Minnesota Press, Minneapolis: 93-124.
- Oldfield, F., J. A. Dearing, R. Thompsom & S. Garret-Jones, 1978. Some magnetic properties of lake sediments and their possible links with erosion rates. Pol. Arch. Hydrobiol. 25: 321-331.
- Oldfield, F., B. A. Maher & P. G. Appleby, 1989. Sediment source variations and lead-210 inventories in recent Potomac Estuary sediment cores. J. Quat. Sci. 4: 189-200.
- Plumb, R. H., 1981. Procedures for Handling and Chemical Analysis of Sediment and Water Samples, Tech. Rep. EPA/CE-81-1. U.S. Army Engineer Waterways Experiment Station, Vicksburg, MS.
- Rabalais, N. N., W. J. Wiseman, Jr. & R. E. Turner, 1994. Comparison of continuous records of near-bottom dissolved oxygen from the hypoxia zone of Louisiana. Estuaries 17: 850-861.
- Rada, R. G., J. G. Wiener, P. A. Bailey & D. E. Powell, 1990. Recent influxes of metals into Lake Pepin, a natural lake on the upper Mississippi River. Arch. Environ. Contam. Toxicol. 19: 712-716.
- Reavie, E. D., R. I. Hall & J. P. Smol, 1995. An expanded weighted-averaging model for phosphorus concentrations from diatom assemblages in eutrophic British Columbia (Canada) lakes. J. Paleolimnology 14: 49-67.
- Rippey, B. & N. J. Anderson, 1996. Reconstruction of lake phosphorus loading and dynamics using the sedimentary record. Environ. Sci. Technol. 30: 1786-1788.
- Schelske, C. L., J. A. Robbins, W. D. Gardner & D. J. Conley, 1988. Sediment record of biogeochemical responses to anthropogenic perturbations of nutrient cycles in Lake Ontario. Can. J. Fish. Aquat. Sci. 45: 1291-1303.
- Smith, R. A., R. B. Alexander & K. J. Lanfear, 1994. Stream Water Quality in the Conterminous United States -- Status and Trends of Selected Indicators during the 1980s; U.S. Geological Survey Water-Supply Paper 2400. U.S. Geological Survey, Denver, CO.
- Smol, J. P., 1985. The ratio of diatom frustules to chrysophycean statospores: a useful paleolimnological index. Hydrobiologia 123: 199-208.
- Smol, J. P., 1992. Paleolimnology: an important tool for effective ecosystem management. J. Aquatic Ecosyst. Health 1: 49-58.
- Smol, J. P., 1993. Paleophycology of a high arctic lake near Cape Herschel, Ellesmere Island. Can. J. Bot. 62: 2195-2204.

- Stark, J. R., W. J. Andrews, J. D. Fallon, A. L. Fong, R. M. Goldstein, P. E. Hanson, S. E. Kroening & K. E. Lee, 1996. Water-Quality Assessment of Part of the Upper Mississippi River Basin, Minnesota and Wisconsin Environmental Setting and Study Design; USGS Water-Resources Investigations Report 96-4098. U.S. Geological Survey, Denver, CO, 62 pp.
- Stoermer, E. F., 1978. Phytoplankton assemblages as indicators of water quality in the Laurentian Great Lakes. Trans. Amer. Microscop. Soc. 97: 2-16.
- Stoermer, E. F., G. Emmert & C. L. Schelske, 1989. Morphological variations of *Stephanodiscus niagarae* Ehrenb. (Bacillariophyta) in a Lake Ontario sediment core. J. Paleolimnology 2: 227-236.
- Stoermer, E. F., J. A. Wolin, C. L. Schelske & D. C. Conley, 1985. Variations in *Melosira islandica* valve morphology related to eutrophication and silica depletion. Limnol. Oceanogr. 30: 414-418.
- Stuiver, M. & G. W. Pearson., 1993. High-precision bidecadal calibration of the radiocarbon time scale, AD 1950-500 BC and 2500-6000 BC. Radiocarbon 35: 1-23.
- Stuiver, M. & P. J. Reimer, 1993. Extended ¹⁴C database and revised CALIB radiocarbon calibration program. Radiocarbon 35: 215-230.
- ter Braak, C. J. F. & I. C. Prentice, 1988. A theory of gradient analysis. Adv. Ecol. Res. 18: 271-317.
- Turner, R. E. & N. N. Rabalais, 1994. Coastal eutrophication near the Mississippi River delta. Nature 368: 619-621.
- Urban, N. R., S. J. Eisenreich, D. F. Grigal & K. T. Schurr, 1990. Mobility and diagenesis of Pb and ²¹⁰Pb in peat. Geochim. Cosmochim. Acta 54: 3329-3346.
- Van Metre, P. C., E. Callender & C. C. Fuller, 1997. Historical trends in organochlorine compounds in river basins identified using sediment cores from reservoirs. Environ. Sci. Technol. 31: 2339-2344.
- Wright, H. E., Jr., 1991. Coring tips. J. Paleolimnology 6: 37-49.
- Wright, H. E., Jr., D. H. Mann & P. H. Glaser, 1983. Piston corers for lake and wetland sediments. Ecology 65: 657-659.
- Wright, H. E., K. Lease & S. Johnson, 1998. Glacial River Warren, Lake Pepin, and the environmental history of southeastern Minnesota. In C. J. Patterson (ed.), Contributions to Quaternary Studies in Minnesota. Minnesota Geological Survey Report of Investigations 49, Minneapolis, MN: 131-140.
- Zumberge, J. H., 1952. The lakes of Minnesota: their origin and classification. Minn. Geological Survey Bull. 35: 58-63.

FIGURES

- Figure 1. Location of core sites, depositional regions, and seismic transects in Lake Pepin. Inset map shows major sub-basins of the upper Mississippi River above Lake Pepin.
- Figure 2. Exemplary seismic traces along core-transect II; black bars mark core locations.
- Figure 3. Alignment of seismic traces with magnetic-susceptibility profiles and loss-on-ignition data for cores I.2 and I.3.
- Figure 4. Magnetic susceptibility profiles for all cores; correlated magnetic features (mf) indicated by connecting lines; cores selected for detailed stratigraphic analysis in bold type; cores selected for top/bottom analysis underlined; cores selected for pollen analysis as noted.
- Figure 5. Loss-on-ignition profiles for 10 detailed cores; sediment depths corresponding to settlement horizon (*Ambrosia* rise) indicated by dashed lines.
- Figure 6. Pollen percentage diagrams showing selected taxa for five detailed cores; approximate depths of settlement horizon (*Ambrosia* rise) and "big-woods" transition shown as dashed and dotted lines, respectively.
- Figure 7. Calibrated AMS ¹⁴C dates and corresponding magnetic features (mf) for cores III.2 and II.1; mf-7 was set at 1830 (time of European settlement), and mf-7.5 and 8 were set at their corresponding ¹⁴C dates. The asymmetry of the error bars results from multiple intercepts of the ¹⁴C calibration curve as modeled by the CALIB program.
- Figure 8. Cesium-137 profiles for detailed (bold) and top/bottom (underlined) cores. Maxima represent 1964 time horizon.
- Figure 9. Total ²¹⁰Pb activity plots for 10 detailed cores. Supported ²¹⁰Pb indicated by dashed line; error bars represent ± 1 s.d.
- Figure 10. Uncorrected ²¹⁰Pb dates calculated by the c.r.s model for individual cores, showing deviations from synchronous stratigraphic markers. Only four cores had dates extending to the 1860 pollen (LOI) marker; error bars represent ± 1 s.d.
- Figure 11. Sediment accumulation rates vs. sediment age for 10 detailed cores.
- Figure 12. Mean grain size vs. sediment age for 10 detailed cores.
- Figure 13. Concentrations for total phosphorus and P fractions vs. sediment age for 10 detailed cores. Plots are stacked; total is sum of NaOH-P, HCl-P, and residual-P.
- Figure 14. Percent of sediment attributed to the Minnesota River based on heavy-mineral elemental chemistry and CMB modeling for 10 detailed cores.
- Figure 15. Average percent of sediment attributed to the Minnesota River before and after European settlement (1830) in 10 detailed cores (a), post-settlement increase in proportion of sediment attributed the Minnesota River (b), and whole-basin sediment loading from the Minnesota and Mississippi/St. Croix river basins.
- Figure 16. Annual whole-basin sediment (a) and phosphorus (b) accumulation in Lake Pepin.

- Figure 17. Diatom concentration diagram from core V.1 showing selected taxa (a). Total-phosphorus reconstructions by weighted-averaging regression and calibration; TP-1 and TP-2 represent alternative reconstructions based on species optima from different datasets (see text) (b). Percentage of total diatoms in each sample used in the reconstruction (c).
- Figure 18. Annual discharge of the Mississippi River at the USGS gauging station at Prescott, Wisconsin (a); solid line indicates period of actual measurement; dashed lines represent inferred flows scaled from older records of the Mississippi River at St. Paul, Minnesota (St.P) and the St. Croix River at St. Croix Falls, Wisconsin (St.C). Mean decadal discharge (b) of the Mississippi River at Prescott.
- Figure 19. Reconstructed phosphorus mass-balance for Lake Pepin; phosphorus loading is calculated as the sum of P outflow and P sedimentation (see text).
- Figure 20. Linear and volumetric sedimentation rates for Lake Pepin (a), projected lake volume from 1990 datum (b), and % volume loss relative to 1830 datum (c).



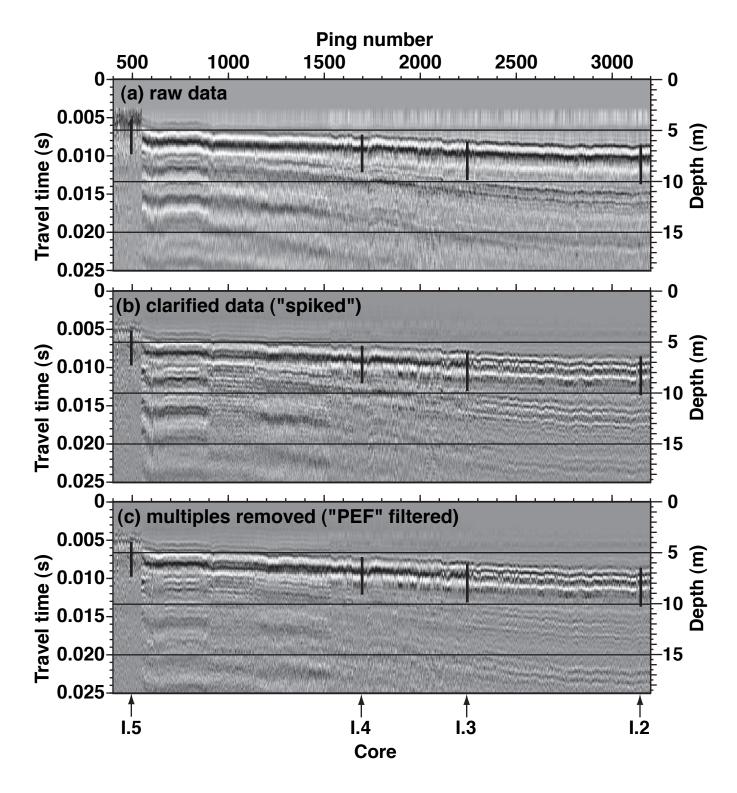


Figure 2. Seismic trace along core transect I, Lake Pepin, 1996

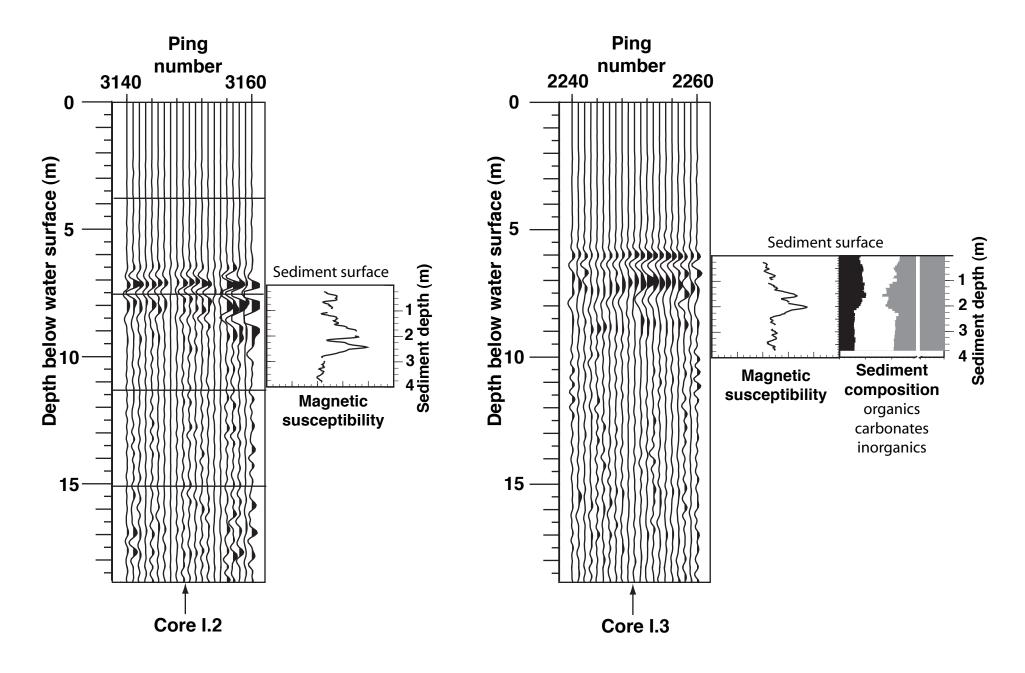


Figure 3. Comparison of seismic profiles and stratigraphic features in cores I.2 and I.3, Lake Pepin

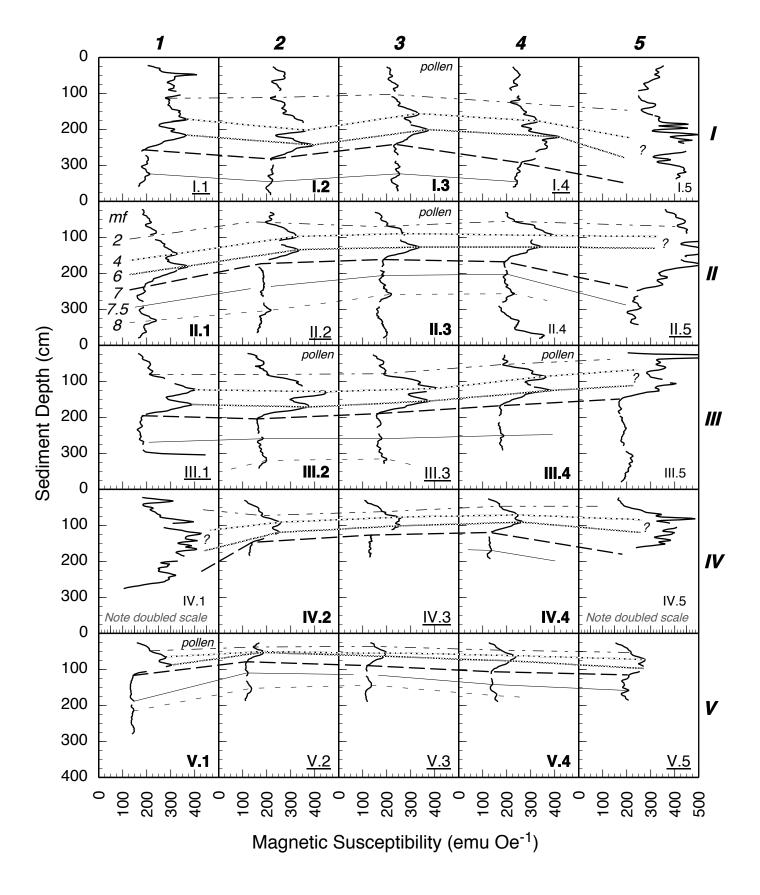


Figure 4. Whole-core magnetic susceptibility profiles

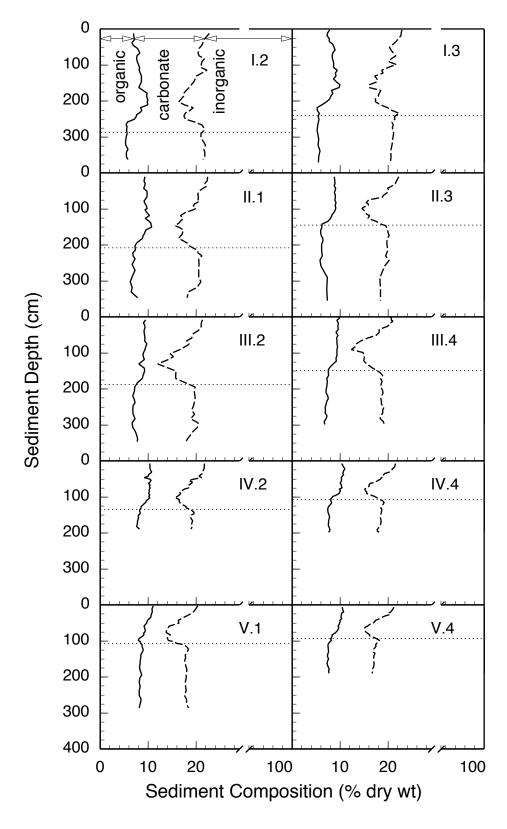
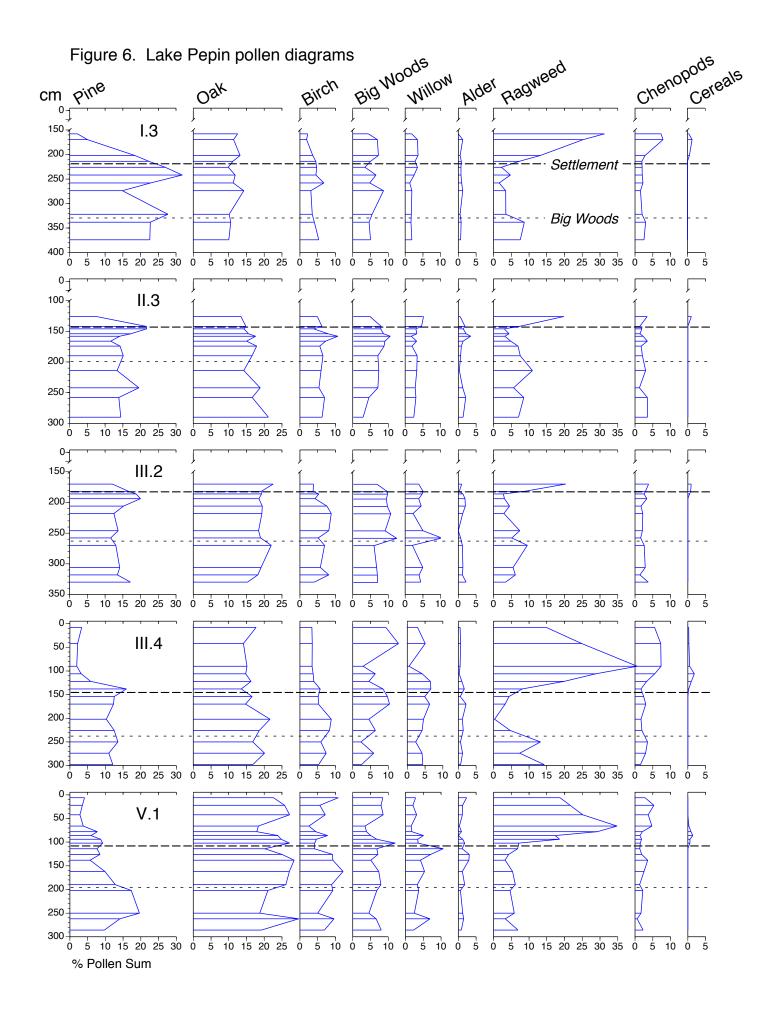


Figure 5. Loss-on-ignition data for Lake Pepin cores



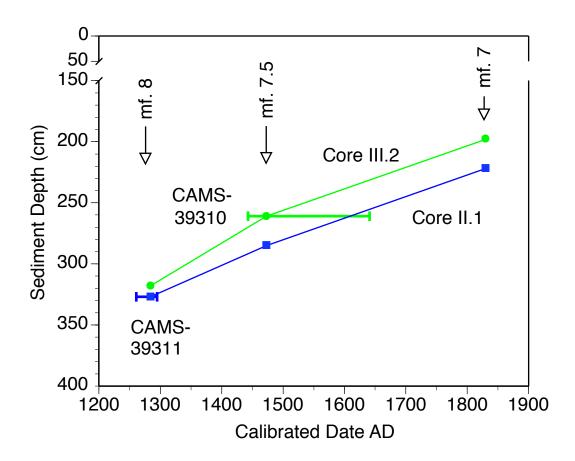


Figure 7. Lake Pepin radiocarbon dates

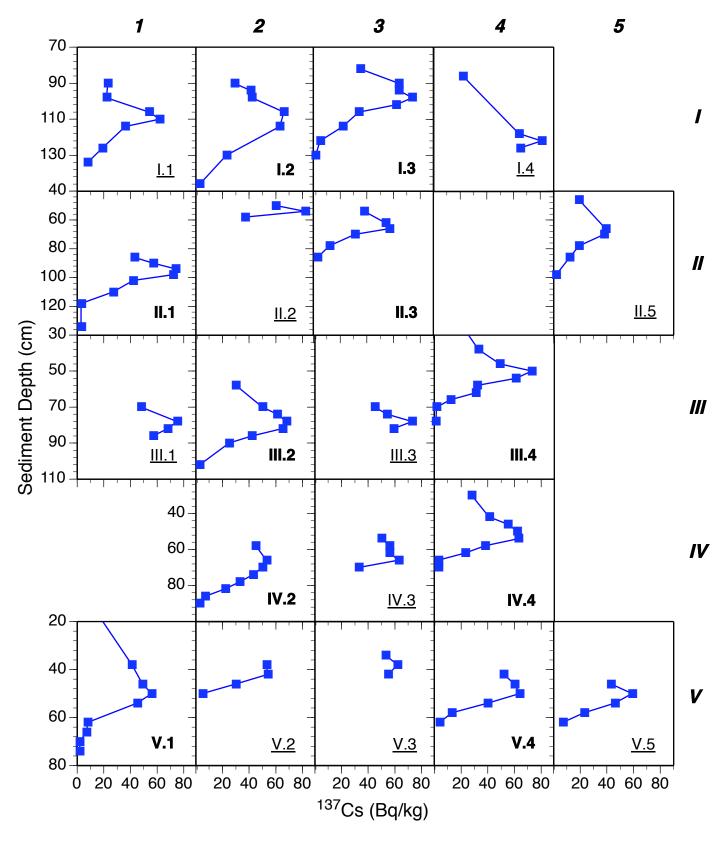
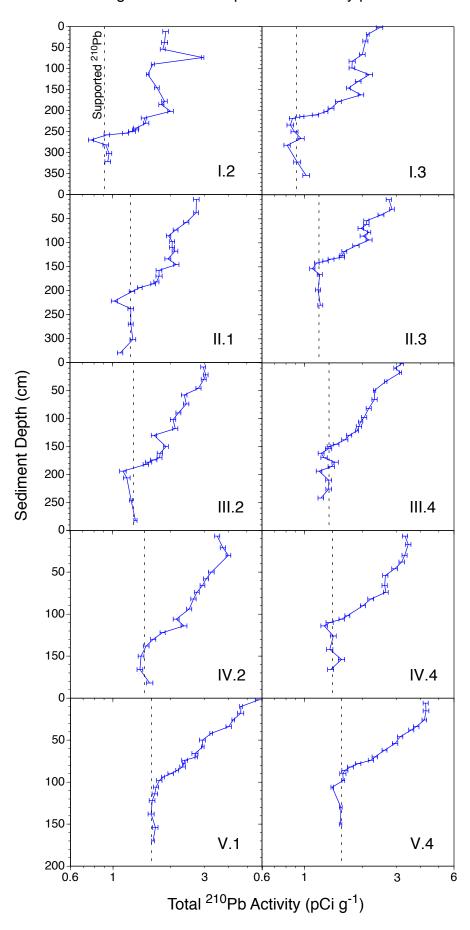
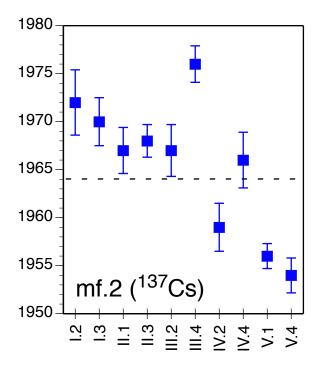
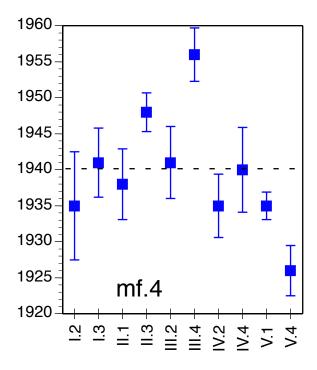


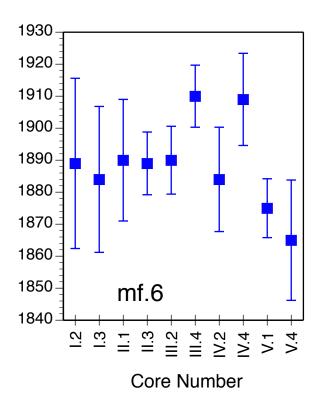
Figure 8. Lake Pepin ¹³⁷Cs profiles

Figure 9. Lake Pepin ²¹⁰Pb activity profiles









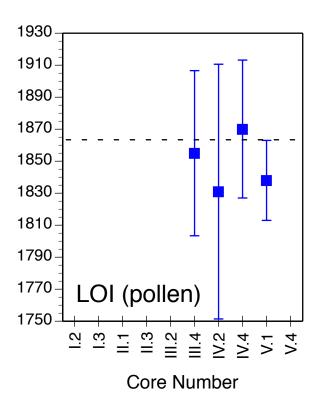


Figure 10. Uncorrected ²¹⁰Pb dates at stratigraphic markers

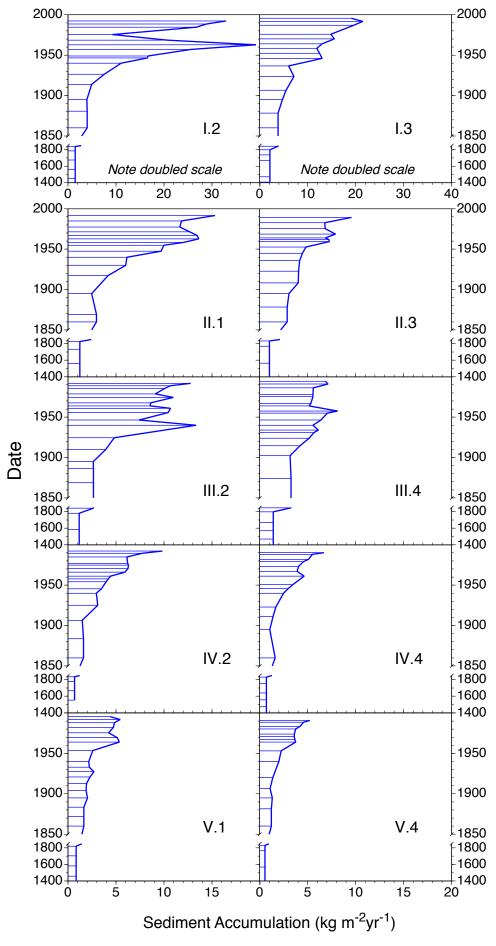


Figure 11. Lake Pepin sediment accumulation rates

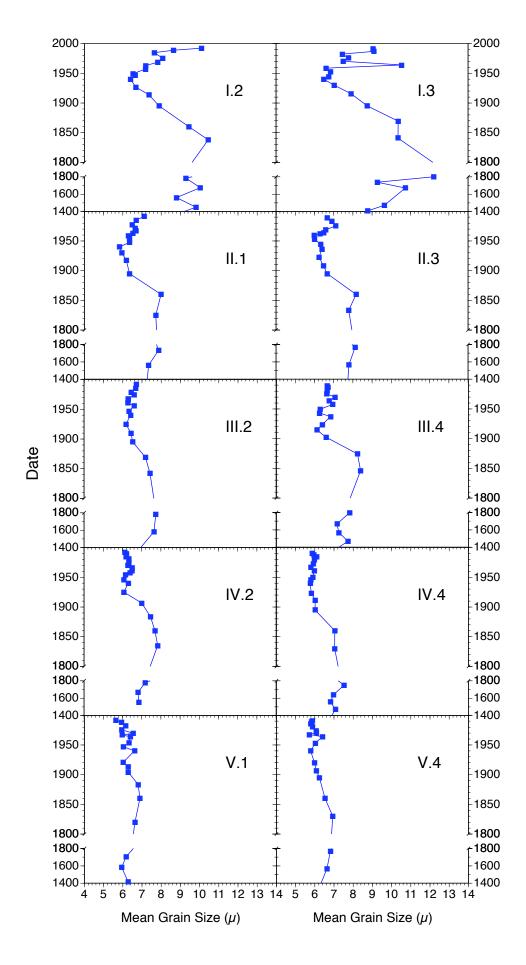


Figure 12. Lake Pepin sediment grain size

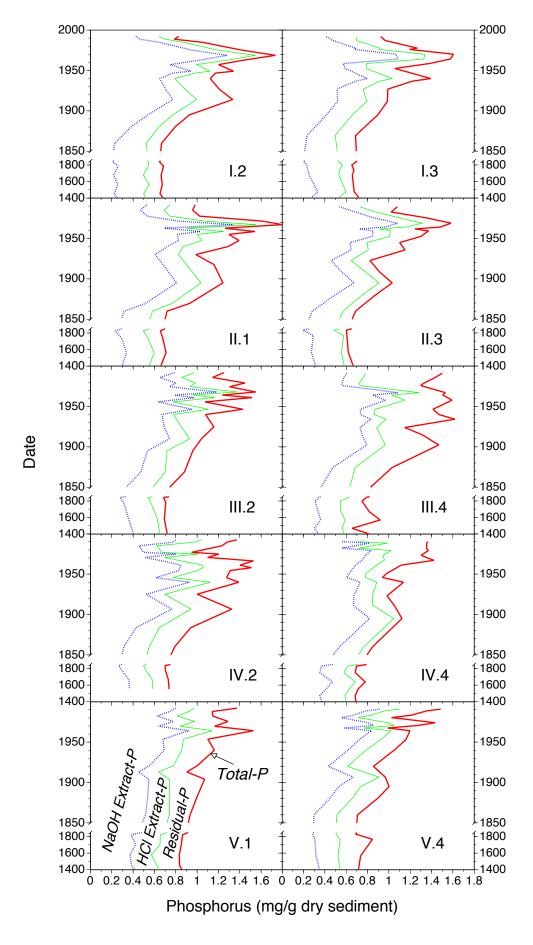


Figure 13. Lake Pepin sediment phosphorus concentration

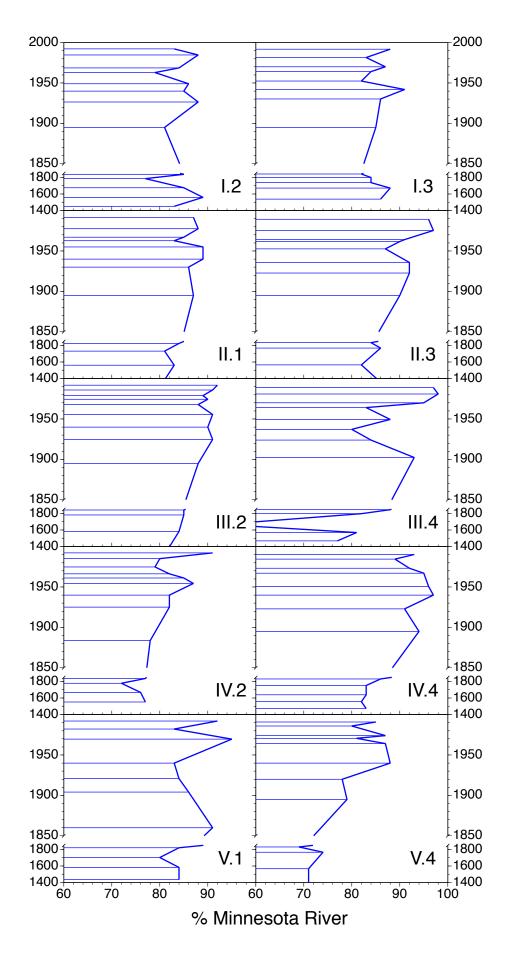


Figure 14. Lake Pepin sediment source apportionment

Figure 15. Relative source contributions and loading of sediment to Lake Pepin

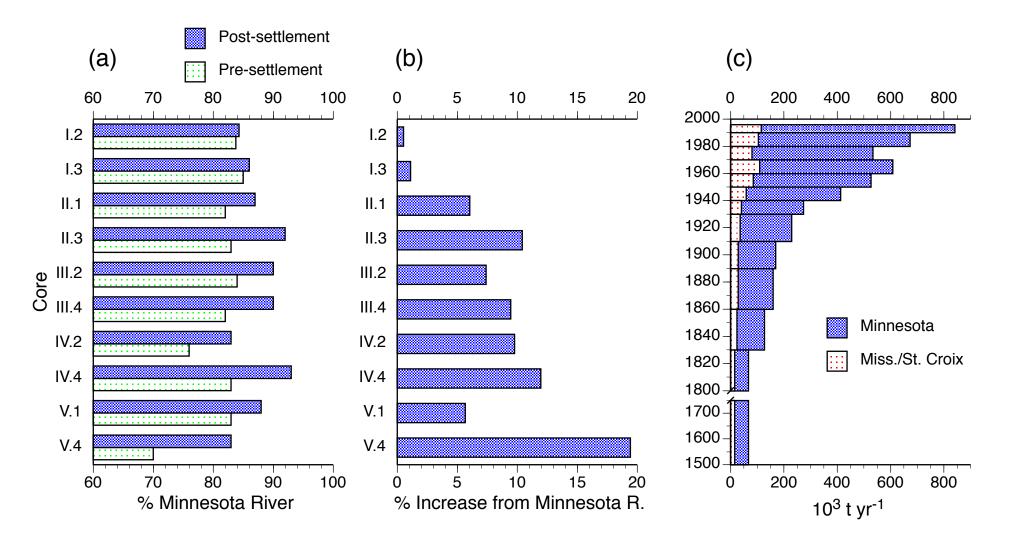
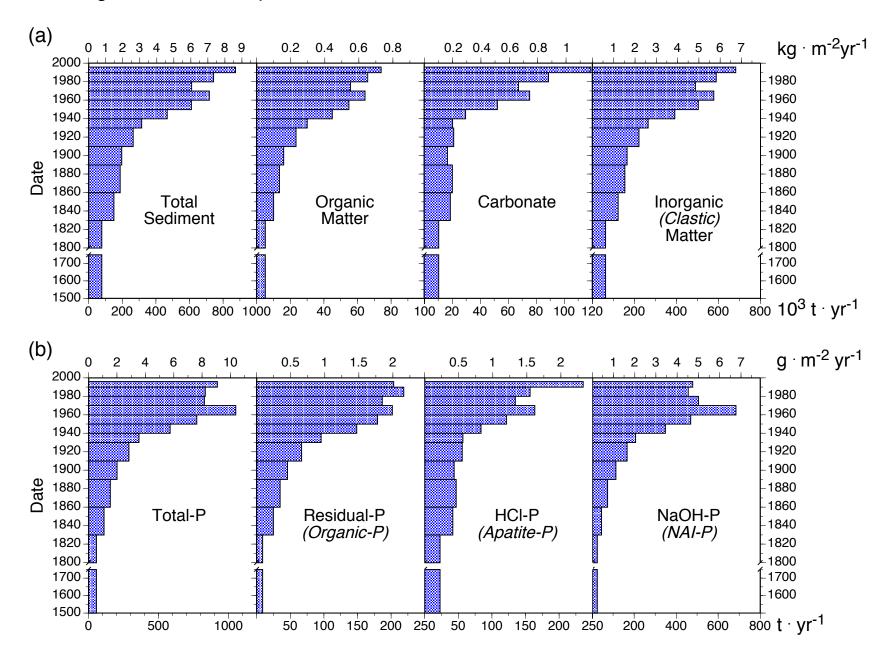
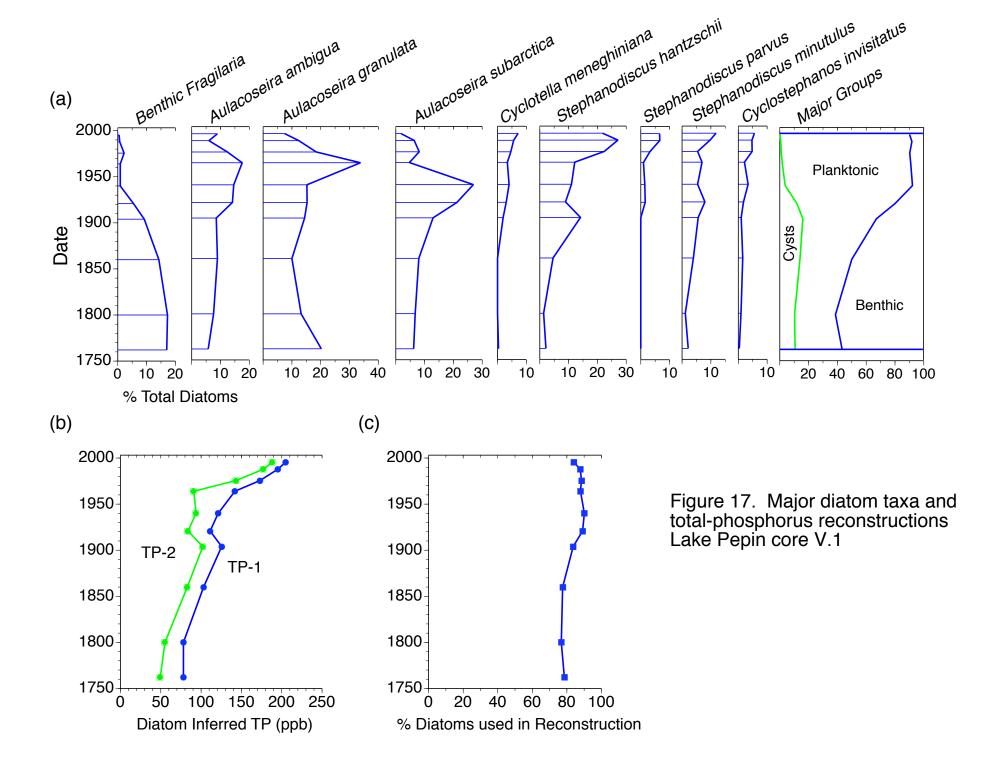


Figure 16. Lake Pepin whole-basin annual sediment accumulation





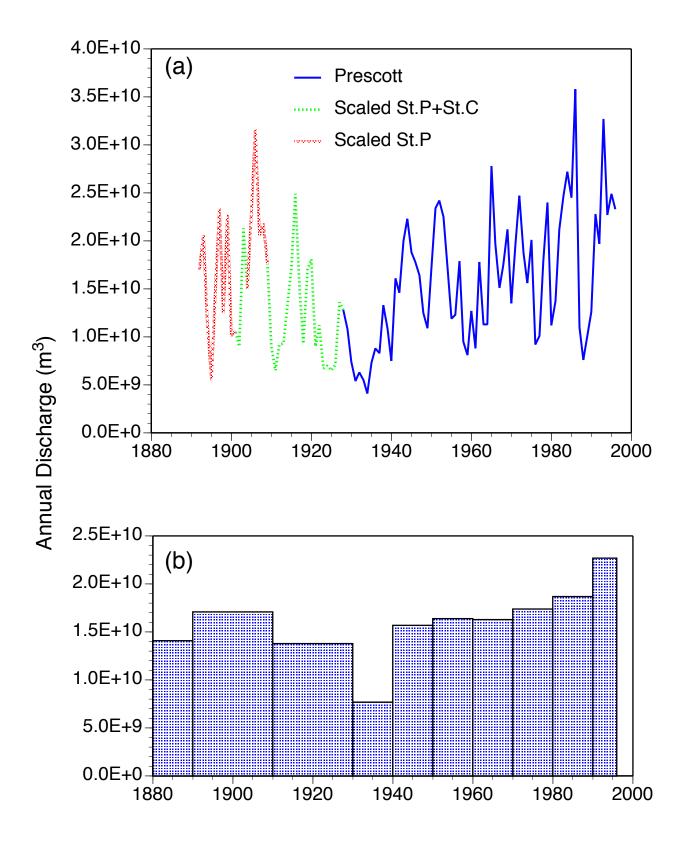


Figure 18. Annual discharge of the Mississippi R. at Prescott WI

Figure 19. Lake Pepin reconstructed phosphorus mass balance

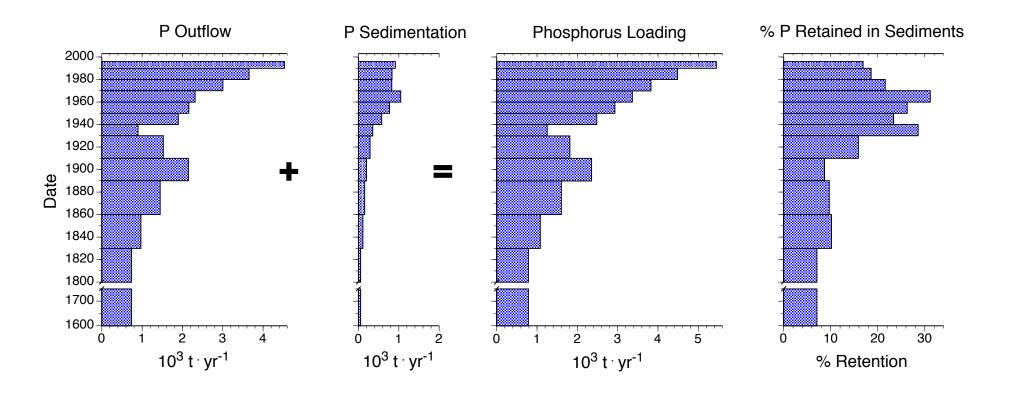


Figure 20. Lake Pepin sediment infilling

